

1 **Hydrodynamic feedbacks of salt-marsh loss in the shallow microtidal back-barrier**  
2 **lagoon of Venice (Italy)**

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18 **Key Points:**

- 19 • Effects of salt-marsh lateral erosion on the characteristics of tides and waves are  
20 investigated for the microtidal Venice Lagoon
- 21 • Salt-marsh loss primarily enhances mean-high water levels and wave heights due to  
22 reduced friction and longer wind fetches
- 23 • Hydrodynamic changes depend on salt-marsh spatial distribution and elevation, as well as  
24 on the characteristics of tidal and wind forcings

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26

27 **Keywords**

28 morphodynamics, salt marshes, salt-marsh erosion, back-barrier system, Venice Lagoon

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30

**Abstract**

31 Extensive loss of salt marshes in back-barrier tidal embayments is ongoing worldwide as  
32 a consequence of land-use changes, wave-driven lateral marsh erosion, and relative sea-level rise  
33 compounded by mineral sediment starvation. However, how salt-marsh loss affects the  
34 hydrodynamics of back-barrier systems and feeds back into their morphodynamic evolution is  
35 still poorly understood. Here we use a depth-averaged numerical hydrodynamic model to  
36 investigate the feedback between salt-marsh erosion and hydrodynamic changes in the Venice  
37 Lagoon, a large microtidal back-barrier system in northeastern Italy. Numerical simulations are  
38 carried out for past morphological configurations of the lagoon dating back up to 1887, as well  
39 as for hypothetical scenarios involving additional marsh erosion relative to the present-day  
40 conditions. The progressive loss of salt marshes significantly impacted the lagoon  
41 hydrodynamics, both directly and indirectly, by amplifying high-tide water levels, reducing  
42 wind-wave energy dissipation, and critically affecting tidal asymmetries across the lagoon.  
43 Restoration projects and manmade protection of marsh margins, which have been implemented  
44 over the past few decades, limited the detrimental effects of marsh loss on the lagoon  
45 hydrodynamics, while not substantially changing the risk of flooding in urban lagoon  
46 settlements. Compared to previous studies, our analyses suggest that the hydrodynamic response  
47 of back-barrier systems to salt-marsh erosion is extremely site-specific, depending closely on the  
48 morphological characteristics of the embayment as well as on the external tidal and wind  
49 forcings.

50

**1 Introduction**

51 Tidal back-barrier lagoons represent critical environments at the interface between  
52 terrestrial, freshwater, and marine habitats (Flemming, 2012; Levin et al., 2001; Pérez-Ruzafa et  
53 al., 2019; Perillo, 1995), and are especially common along the World's coasts (Boothroyd et al.,  
54 1985; Fitzgerald & Hughes, 2019; Kjerfve, 1994; Stutz & Pilkey, 2011). They consist of  
55 sheltered embayments separated from the ocean by a system of barrier islands (Hesp, 2016)  
56 interrupted by tidal inlets (De Swart & Zimmerman, 2009). Back-barrier lagoons support high  
57 biodiversity, densely populated urban settlements, and florid economies (Barbier et al., 2011;  
58 Costanza et al., 1997; D'Alpaos & D'Alpaos, 2021). However, accelerating sea-level rise,  
59 reduced sediment supply, enhanced storminess, and increasing anthropogenic pressures  
60 exacerbate the threat to back-barrier lagoons and the communities relying on them (Gilby et al.,  
61 2021; Passeri et al., 2020). Although the current paradigm indicates that future coastal hazards  
62 will be mostly dictated by rising sea levels (Finkelstein & Ferland, 1987; González-Villanueva et  
63 al., 2015), previous studies demonstrated how geomorphological changes in tidal embayments  
64 can feedback into coastal hydrodynamics and enhance coastal hazards (Carniello et al., 2009;  
65 Ferrarin et al., 2015; Orton et al., 2020; Pollard et al., 2019; Ralston et al., 2019; Zhou et al.,  
66 2014). Therefore, investigating the feedback between geomorphological and hydrodynamic  
67 changes in back-barrier systems is of utmost importance to provide reliable assessments of  
68 coastal hazards (Carniello et al., 2009; Donatelli et al., 2018; Donatelli, Kalra, et al., 2020;  
69 Donatelli, Zhang, et al., 2020; Ferrarin et al., 2015; Vinet & Zhedanov, 2011; Zarzuelo et al.,  
70 2018).

71

72 Among the geomorphological features that characterize shallow back-barrier tidal  
73 embayments, salt marshes are especially common and provide a wide number of precious  
ecosystem services, including blue-carbon sequestration (Chmura et al., 2003), environmental

74 remediation (Nelson & Zavaleta, 2012), shoreline protection (Möller et al., 2014; Temmerman et al., 2013), and habitat provision (Hopkinson et al., 2018; Pennings & He, 2021). The alarming rates of salt-marsh loss observed worldwide (Mcowen et al., 2017; Valiela et al., 2009) have prompted extensive studies on salt-marsh ecomorphodynamics (A. D'Alpaos et al., 2007; Fagherazzi et al., 2012; Finotello, D'Alpaos, et al., 2022), as well as on the response of these ecosystems to changing hydrodynamic forcings and inorganic sediment supply (A. D'Alpaos et al., 2007; Finotello et al., 2020; Fitzgerald & Hughes, 2019; Gourgue et al., 2022; Hughes et al., 2021; Mariotti, 2020; Tommasini et al., 2019). (A. D'Alpaos et al., 2007; Finotello et al., 2020; Fitzgerald & Hughes, 2019; Gourgue et al., 2022; Hughes et al., 2021; Mariotti, 2020; Tommasini et al., 2019). In contrast, how salt-marsh loss affects hydrodynamics and the related morphodynamic evolution in shallow coastal bays still remains unclear. This uncertainty is mostly due to the paucity of study cases analyzed so far, thus calling for new insights into the feedback between salt-marsh loss and hydrodynamic changes in shallow back-barrier tidal systems (Donatelli et al., 2018; Donatelli, Zhang, et al., 2020; Silvestri et al., 2018; Vinet & Zhedanov, 2011).

89 Here we focus on the microtidal Venice Lagoon (Italy), where extensive marsh losses have been documented over the last two centuries (Carniello et al., 2009; L. D'Alpaos, 2010; Tommasini et al., 2019). We will pay particular attention to the feedback between wind-wave-induced salt-marsh erosion and changes in the lagoon hydrodynamics. The latter will be investigated using a custom-built, depth-averaged numerical model applied to several past morphological configurations of the Lagoon, each reconstructed based on available historical topographic and bathymetric maps. Moreover, exploratory simulations will be performed to unravel the hydrodynamic consequences of the failure to restore and protect salt marshes over the past few decades, which would have led to a further reduction in the total marsh extent compared to the present-day conditions.

## 99 2 Geomorphological Setting

100 Located in the northern Adriatic Sea, and characterized by an area of 550 km<sup>2</sup>, the Venice Lagoon is the largest brackish waterbody in the Mediterranean Basin. The Lagoon formed over the last 7500 years covering alluvial Late Pleistocene, silty-clayey deposits locally known as *Caranto* (Zecchin et al., 2008). The Lagoon's present-day morphology is characterized by the presence of three inlets, namely, from North to South: Lido, Malamocco, and Chioggia (Figure 1a-d). Tides follow a semidiurnal microtidal regime, with a mean spring tidal range of 1 m and maximum tidal oscillations of about 0.75 m around Mean Sea Level (MSL) (e.g., D'Alpaos et al. 2013; Valle-Levinson et al. 2021).

108 Meteorological surges often overlap astronomical tides, thus producing significantly high (low) tides when atmospheric pressure is low (high). In addition, wind-related processes are critical for both the hydrodynamics and morphodynamics of the lagoon, with seasonal wind-storm events exerting a prominent morphodynamic control over decadal to centenary timescales (see Carniello et al. 2009, 2012). Detailed analyses of the wind climate show negligible interannual variability of wind energy (see Text S1 and Figure S1 in the Supporting Information). The most morphologically and hydrodynamically meaningful wind-storm events are those associated with *Bora* and *Sirocco* winds (Figure 1e). The north-easterly *Bora* winds blow almost parallel to the major axis of the lagoon, thus producing pronounced water-level setups in the southern lagoon and generating large waves (significant wave height  $H_s > 1$  m). Such

118 waves promote significant resuspension of sediments from the tidal mudflats. In contrast,  
119 *Sirocco* winds blow from the South-East and cause large water-level setups in the northern  
120 Adriatic Sea, often leading to extensive flooding of Venice city and other settlements within the  
121 lagoon.

122 Over the last centuries, the hydrodynamics of the lagoon was severely affected by  
123 anthropogenic interventions (L. D'Alpaos, 2010; Ferrarin et al., 2015). First, by the end of the  
124 16<sup>th</sup> century, all the major rivers debouching into the lagoon were diverted into the open sea, thus  
125 almost completely eliminating fluvial sediment input. Second, between the 1900s and 1970s,  
126 extensive land reclamation projects were carried out, especially along the landward margin of the  
127 lagoon, thus importantly reducing the total area open to the propagation of tides (see Figure 2).  
128 During the same period, the extraction of groundwater and natural gas for industrial purposes  
129 significantly accelerated the natural rates of soil subsidence (Carbognin et al., 2004; Gatto &  
130 Carbognin, 1981; Zanchettin et al., 2021). Moreover, in order to allow for increasingly bigger  
131 ships to cruise within the lagoon, two large waterways, namely the Vittorio Emanuele and the  
132 Malamocco-Marghera channels (Figure 2c,d), were excavated in 1925 and 1968, respectively.  
133 Major changes in the lagoon hydrodynamics were due to the construction of jetties at the lagoon  
134 inlets aimed to ensure water depths suitable for commercial ship traffic (Figure 1b-d and Figure  
135 2a-d). The jetties at the Malamocco inlet were constructed between 1839 and 1872, whereas at  
136 the Lido inlet the northern jetty was completed in 1887 (see Figure 2a), with the southern jetty  
137 added later in 1892 (see Figure 2b). Finally, the jetties at the Chioggia inlet were built between  
138 1910 and 1934 (Figure 2c). On the one hand, the jetties reduced the width of the inlets, thus  
139 resulting in considerable deepening as foreseen during the design phase (Figure 2a-c). On the  
140 other hand, they caused critical changes in the lagoon hydro- and morpho-dynamic regimes.  
141 Since the construction of the jetties, changes in the tidal regime within the lagoon have been  
142 much more sustained than the typical periodic, multi-annual variations induced by the nodal  
143 modulation of tides in the Adriatic Sea, which are in the order of 4% of the characteristic tidal  
144 range (Amos et al., 2010; Valle-Levinson et al., 2021). Between 1909 and 1973, the tidal range  
145 within the lagoon increased as much as 25% on average (L. D'Alpaos, 2010; Ferrarin et al.,  
146 2015; Tomasin, 1974), with local changes that can be even more pronounced (Finotello et al.,  
147 2019; Finotello, Capperucci, et al., 2022; Silvestri et al., 2018).

148 Changes in the lagoon hydrodynamics due to the construction of the jetties, coupled with  
149 eustatic sea-level rise (average value  $1.23 \pm 0.13$  mm/year between 1872 and 2019;  $2.76 \pm 1.75$   
150 mm/year between 1993 and 2019; see Zanchettin et al. 2021), critically impacted on the lagoon  
151 morphological evolution, triggering positive morphodynamic feedbacks. Progressively larger  
152 portions of the lagoon became ebb-dominated, especially close to the inlets where the jetties  
153 produced strong flow asymmetries. Asymmetric tidal flows enhanced the export of fine  
154 sediments and prevented the import of sediment carried in suspension by longshore currents (L.  
155 D'Alpaos, 2010). This condition, worsened by anthropogenically-induced starvation of fluvial  
156 sediments, set a negative sediment budget and resulted in a generalized loss of salt marshes  
157 (Carniello et al., 2009; L. D'Alpaos, 2010; Tommasini et al., 2019; see Figure 2a-f,k). Reduced  
158 marsh coverage lengthened wind fetches, thus favoring the formation of higher, more energetic  
159 waves, which further enhanced lateral marsh retreat a prompted erosion of tidal mudflats  
160 (Carniello et al., 2009; Finotello et al., 2020; Leonardi, Ganju, et al., 2016; Marani, D'Alpaos, et  
161 al., 2011; Mariotti & Fagherazzi, 2013; Tommasini et al., 2019; see Figure 2). Mudflat  
162 deepening (Figure 2l), exacerbated by eustatic sea-level rise and both natural and anthropogenic-

163 induced subsidence, promoted the formation of even higher wind waves, which in turn favored  
 164 additional erosion of salt marshes and mudflats through a positive feedback loop.

165 Further manmade modifications of the inlet morphologies were carried out between 2006  
 166 and 2014 to accommodate the mobile floodgates of the Mo.S.E. (acronym for “Modulo  
 167 Sperimentale Elettromeccanico”, Electromechanical Experimental Module) system (Figure  
 168 1b,c,d), designed to protect the city of Venice and other lagoon settlements from extensive  
 169 floodings (Mel, Viero, et al., 2021). These interventions slightly increased hydraulic resistances  
 170 at the inlets and led to both a reduction of tidal amplitudes and an increase in tidal-phase delays  
 171 within the lagoon (Ghezzi et al., 2010; Matticchio et al., 2017). Salt-marsh erosion is still  
 172 ongoing nowadays, though at much lower rates compared to the last century (Figure 2k). This is  
 173 due to a series of critical interventions, aimed at safeguarding and restoring salt marshes, that  
 174 have been put in place by the Venice Water Authority since the early 1990s, with additional  
 175 more recent contributions by some EU-funded LIFE projects (Barausse et al., 2015; Tagliapietra  
 176 et al., 2018; Tommasini et al., 2019; see also [www.lifevimine.eu/](http://www.lifevimine.eu/) and  
 177 [www.lifelagoonrefresh.eu/](http://www.lifelagoonrefresh.eu/)). At present, about 12 % of the existing salt marshes are either  
 178 entirely artificial or at least partially restored (see purple lines in Figure 1a), and a good portion  
 179 of the remaining natural ones are protected against lateral erosion by manmade wood piling or  
 180 berms (see Figure S2 in the Supporting Information). Clearly, without these restoration and  
 181 conservation efforts, the total area of salt marshes would be significantly smaller than it currently  
 182 is.

183 Finally, it is worthwhile noting that operations of the Mo.S.E. floodgates will further reduce the  
 184 resilience of salt marshes to rising relative sea levels by reducing inorganic deposition during  
 185 storm-surge events which, though episodic, critically contribute to marsh vertical accretion  
 186 (Tognin et al., 2021, 2022).

## 187 **3 Methods**

### 188 **3.1 Numerical model**

189 We employed the bidimensional, depth-averaged, finite element numerical model  
 190 developed by Carniello et al. (2005, 2011), which is suitable to reproduce the hydrodynamics of  
 191 shallow tidal basins driven by tidal flows and wind fields. In the following, we report a brief  
 192 description of the model and refer the reader to Carniello et al. (2005, 2011) for further details.  
 193 The model consists of two coupled modules, namely a hydrodynamic and a wind-wave module,  
 194 and will be referred to as WWTM (Wind-Wave Tidal Model) hereinafter.

195 The hydrodynamic model solves a suitably-modified version of the shallow water  
 196 equations (Defina 2000) to account for wetting and drying processes in very shallow and  
 197 irregular domains:

198

$$199 \quad \frac{\partial q_x}{\partial t} + \frac{\partial}{\partial x} \left( \frac{q_x^2}{Y} \right) + \frac{\partial}{\partial y} \left( \frac{q_x q_y}{Y} \right) - \left( \frac{\partial R_{xx}}{\partial x} + \frac{\partial R_{xy}}{\partial y} \right) + \frac{\tau_{bx}}{\rho} - \frac{\tau_{wx}}{\rho} + gY \frac{\partial H}{\partial x} = 0$$

$$200 \quad \frac{\partial q_y}{\partial t} + \frac{\partial}{\partial x} \left( \frac{q_x q_y}{Y} \right) + \frac{\partial}{\partial y} \left( \frac{q_y^2}{Y} \right) - \left( \frac{\partial R_{xy}}{\partial x} + \frac{\partial R_{yy}}{\partial y} \right) + \frac{\tau_{by}}{\rho} - \frac{\tau_{wy}}{\rho} + gY \frac{\partial H}{\partial y} = 0$$

201

$$\eta \frac{\partial H}{\partial t} + \frac{\partial q_x}{\partial x} + \frac{\partial q_y}{\partial y} = 0$$

202 where  $t$  is time, the  $x$  and  $y$  subscripts denote the directions of a given variable in a Cartesian  
 203 reference system,  $q$  is the flow rate per unit width,  $R$  represents the depth-averaged Reynolds  
 204 stresses,  $\tau_b$  is the bottom shear stress produced by tidal currents,  $\tau_w$  is the wind-induced shear  
 205 stress at the free surface whose elevation is  $H$ ,  $\rho$  stands for water density,  $g$  is the gravitational  
 206 acceleration,  $Y$  is the water volume per unit area ponding the bottom (i.e., the equivalent water  
 207 depth) and  $\eta$  is the wet fraction of the computational domain which accounts for surface  
 208 irregularities during the wetting and drying processes (see Defina, 2000, for a detailed  
 209 description of the hydrodynamic equations).

210 The hydrodynamic module provides the wind-wave module with water levels and depth-  
 211 averaged velocities that are employed for calculating wave group celerity and bottom shear  
 212 stresses (induced by both wind waves and tidal currents), as well as for evaluating the influence  
 213 of flow depth on wind-wave propagation.

214 The wind-wave module employs the same computational grid of the hydrodynamic  
 215 model to solve the wave action conservation equation (Holthuijsen et al., 1989). The latter is  
 216 simplified by assuming that the direction of wave propagation instantaneously readjusts to match  
 217 the wind direction (i.e., neglecting refraction). The module describes the evolution of the wave  
 218 action density ( $N_0$ ) in the frequency domain, which reads (Carniello et al., 2011):

219

220

$$\frac{\partial N_0}{\partial t} + \frac{\partial}{\partial x} c'_{gx} N_0 + \frac{\partial}{\partial y} c'_{gy} N_0 = S_0$$

221

222 where  $c'_{gx}$  and  $c'_{gy}$  represent the wave group celerity in the  $x$  and  $y$  direction, respectively, and  
 223 are used to approximate the propagation speed of  $N_0$  (Carniello et al., 2005; Holthuijsen et al.,  
 224 1989), while  $S_0$  represents all the source terms describing the external phenomena contributing to  
 225 wave energy variations, which can be either positive (wind energy input) or negative (bottom  
 226 friction, white capping, and depth-induced breaking). Based on the relationship between peak-  
 227 wave period and local wind speed and water depth (Young & Verhagen, 1996), the model can  
 228 compute both the spatial and temporal distribution of the wave period. Linear wave theory also  
 229 allows one to relate the local significant wave height to wind-wave-induced bottom shear  
 230 stresses ( $\tau_{ww}$ ). The nonlinear interactions between  $\tau_{ww}$  and current-induced bottom shear  
 231 stresses ( $\tau_b$ ) are accounted for through the empirical Soulsby's (1995) formulation, which  
 232 enhances the value of the total bottom shear stress ( $\tau_{wc}$ ) beyond the mere sum of  $\tau_b$  and  $\tau_{ww}$   
 233 (see details in Carniello et al., 2005, their equations (26) and (27)).

234

## 3.2 Numerical simulations

235

### 3.2.1 Computational Grids

236

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238

Numerical simulations were performed considering ten different morphological  
 configurations of the Venice Lagoon (Figure 2). Six of these configurations represent past-  
 lagoon morphologies reconstructed from available topographic and bathymetric data (Figure 2a-

239 f), whereas the additional four configurations consist of hypothetical scenarios characterized by  
240 additional marsh loss relative to the present-day lagoon morphology (Figure 2g-j). Specifically,  
241 to understand how marsh loss affected the hydrodynamics of the Venice Lagoon in the past, we  
242 utilized six already existing WWTM computational grids representing the morphological  
243 configurations of the lagoon from 1887 to 2014 (Figure 2). Each grid faithfully reproduces the  
244 lagoon morphological features at the time of the selected topobathymetric surveys (see Figure S3  
245 in the Supporting Information). The 1887 and 1901 grids were constructed based on  
246 “Topographic/hydrographic map of the Venice Lagoon” produced by the Genio Civile of  
247 Venezia in 1901, and are identical to each other except for the different morphology of the Lido  
248 inlet, where only the northern jetty was present in 1887 while both the jetties were completed in  
249 1901. In contrast, different topographic surveys carried out in 1932, 1970, and 2003 by the  
250 Venice Water Authority (Magistrato alle Acque di Venezia) were employed to create the  
251 computational grids relative to the years 1932, 1970, 2003, and 2014, respectively (Carniello et  
252 al., 2009) (Figure 2a-e). The 2014 computational grid is based on the most recent, 2003,  
253 bathymetric survey and accounts also for the anthropogenic modifications at the three inlets  
254 related to the Mo.S.E. system, which were completed in 2014 (Figure 2f).  
255 The interested reader is referred to Carniello et al. (2005, 2011) and Tognin et al. (2022) for  
256 extensive details regarding the calibration of both the hydrodynamic and the wind-wave model  
257 with applications to the Venice Lagoon. Calibration and testing of the model obviously refer to  
258 the most recent configurations of the lagoon, for which field data are available. In contrast, local  
259 values of the bed-friction coefficient for the older configurations of the lagoon (1932, 1901,  
260 1887) were assumed in analogy with those selected for the calibrated grids, that is, as a function  
261 of local sediment grainsize, bed elevation, and the possible presence of vegetation (e.g., in salt  
262 marshes). For the sake of brevity, we will report here only a summary of the model performances  
263 quantified through the standard Nash-Sutcliffe Model Efficiency (NSE) parameter derived from  
264 previous applications of the model, for which measured field data were available in terms of tidal  
265 levels and significant wave heights within the Lagoon, as well as flow rates at the lagoon inlets  
266 (Carniello et al., 2005, 2011; Tognin et al., 2022). Following the categorization of model  
267 performance proposed by Allen et al. (2007), who considered four categories from excellent to  
268 poor (i.e.,  $NSE > 0.65$  excellent;  $0.5 < NSE \leq 0.65$  very good;  $0.2 < NSE \leq 0.5$  good;  $NSE \leq 0.2$  poor),  
269 our WWTM model is excellent in reproducing tidal levels ( $NSE_{\text{mean}}=0.970$ ,  $NSE_{\text{median}}=0.984$ ,  
270  $NSE_{\text{std}}=0.040$ ), very good to excellent in reproducing significant wave heights ( $NSE_{\text{mean}}=0.627$ ,  
271  $NSE_{\text{median}}=0.756$ ,  $NSE_{\text{std}}=0.357$ ), and excellent in replicating flow rates at the inlets  
272 ( $NSE_{\text{mean}}=0.853$ ,  $NSE_{\text{median}}=0.184$ ,  $NSE_{\text{std}}=0.931$ ). All the statistics were derived based on data  
273 reported in Tognin et al., (2022; their Table S2), as well as in Carniello et al. (2011, their Tables  
274 1,2, and 3).

275 In addition to the above, we also investigated the hydrodynamic effects of additional  
276 marsh losses based on four possible scenarios characterized by different degrees of wind-wave-  
277 driven salt-marsh lateral erosion. These scenarios should be properly interpreted as hypothetical  
278 configurations that the present-day lagoon might have assumed if marsh erosion had been more  
279 intense and/or if - more realistically - the interventions aimed at safeguarding and restoring salt  
280 marshes had not been put in place (Barausse et al., 2015; Tagliapietra et al., 2018; see Figure S2  
281 in the Supporting Information). Thus, our analysis investigates what the effects of not protecting  
282 and restoring the salt marshes might have been, and somehow provides a baseline for assessing  
283 whether or not it is worthwhile to continue such conservation efforts. In the four analyzed  
284 scenarios, we accounted for an overall marsh-area loss equal to 25% (E25, Figure 2g), 50%

285 (E50, Figure 2h), 75% (E75, Figure 2i), and 100% (E100, Figure 2j) relative to the present-day  
286 marsh extent. The computational grids for each of these four scenarios were constructed utilizing  
287 the 2014 computational grid as a baseline, and gradually removing marsh areas following an  
288 approach similar to that proposed by Donatelli et al. (2018). Specifically, computational  
289 elements located along eroding marsh margins were selected and their characteristics in terms of  
290 elevation and roughness were modified to match those of the surrounding tidal flats. Differently  
291 from Donatelli et al. (2018) though, the erosion of salt marshes was not spatially uniform, but  
292 rather occurred proportionally to the local mean power of wind waves striking the marsh edge  
293 (Figure S4 in the Supporting Information). Since previous studies have demonstrated a linear  
294 functional dependence between marsh lateral retreat and the power of incoming waves  
295 (Leonardi, Ganju, et al., 2016; Marani, D'alpaos, et al., 2011; Mel et al., 2022), cell conversion  
296 from salt marsh to mudflat was imposed proportionally to the value of local wind-wave power  
297 derived from literature data (Figure S4 in the Supporting Information; Tommasini et al. 2019;  
298 Finotello et al. 2020).

299 Although salt-marsh loss alters the water volume exchanged between the sea and the lagoon (i.e.,  
300 tidal prism) and the inlet cross-sectional area should adjust consequently according to the  
301 O'Brien-Jarret-Marchi law (A. D'Alpaos et al., 2009; Jarrett, 1976), in the hypothetical scenarios  
302 involving additional marsh erosion the geometry of the lagoon inlets was kept unaltered  
303 compared to the present-day configuration. This is because the current geometry of the inlets is  
304 fixed both horizontally and vertically by the presence of the jetties and the concrete housing  
305 structures built to host the Mo.S.E. floodgates. Besides, the scour processes induced by the jetties  
306 during the last century deepened the inlets down to the overconsolidated *Caranto* layer, which  
307 would have prevented any further deepening even if the Mo.S.E. barriers had not been  
308 constructed.

### 309 *3.2.2 Boundary Conditions*

310 In the numerical model, water levels are imposed at the seaward boundary of the  
311 computational domain representing the portion of the northern Adriatic Sea in front of the  
312 Venice Lagoon (see Figure 3b). Water-level data are measured at the CNR Oceanographic  
313 Platform, which is located in the Adriatic Sea approximately 15 km away from the coastline.  
314 Because water levels and bed elevations in each computational grid refer to the mean sea level at  
315 the time of each survey, historical rises in relative sea level are implicitly accounted for. Wind  
316 speeds and directions are measured at the "Chioggia Diga Sud" anemometric station (Figure 3b)  
317 and applied to the whole lagoonal basin (see Carniello et al., 2005 for details).

318 All the simulations were carried out employing the same boundary conditions, thus  
319 allowing for direct comparisons between different lagoon configurations. Specifically, the model  
320 was forced using hourly water levels and both wind velocities and directions measured from  
321 November 16<sup>th</sup>, 2005 to December 17<sup>th</sup>, 2005 (Figure 3a). The selected 30-day period is  
322 representative of hydro-meteorological conditions experienced every year by the Venice Lagoon  
323 between October 1<sup>st</sup> and January 30<sup>th</sup>, which is the period typically characterized by the most  
324 significant storm-surge events. More in detail, the cumulative frequency of water levels for the  
325 selected study period is the closest to the average distribution observed between 2000 and 2020  
326 (Figure 3c). Moreover, the study period is characterized by two relatively strong *Bora* wind  
327 events (Figure 3a) that are typical of the wind climate observed in Venice (Figure 3d). Thus, the

328 selected study period allows for focusing both on characteristic tides as well as on representative  
329 storm events.

## 330 **4 Results**

331 The hydrodynamic effects of morphological changes at the whole lagoon scale were  
332 investigated by considering different parameters related to both tides and wind waves.

### 333 **4.1 Water Levels**

334 Concerning water levels, we first focused on the mean tidal range (*MTR*, Figure 4),  
335 computed as the average difference between consecutive high- and low-tide water levels. Results  
336 show a continued increase of *MTR* from 1887 to 2003, though spatially-explicit representations  
337 suggest that this increase is not homogeneous. Increases in *MTR* between 1887 and 1901 are  
338 mostly limited to the northern lagoon (Figure 4a,b), whereas between 1901 and 1932 enhanced  
339 *MTR* values are observed especially in the southern lagoon and in the surroundings of Venice  
340 City (Figure 4b,c). The most pronounced and generalized increase in *MTR* is observed between  
341 1932 and 1970 (Figure 4d), as a result of extensive losses of marshlands, disappearance of minor  
342 branches of tidal channel networks, and generalized tidal-flat deepening (Figure 2d). In contrast,  
343 only minor changes are observed from 1970 onwards (Figure 4d,e,f), with probability  
344 distributions suggesting only a slight increase in *MTR* between 1970 and 2003 followed by a  
345 reduction between 2003 and 2014 (Figure 4k). Numerical simulations involving additional loss  
346 of salt-marsh areas suggest slight reductions in *MTR* proportionally to the percentage of marsh  
347 area being lost (Figure 4l).

348 Since *MTR* does not embed information regarding modifications of high water levels,  
349 changes in the Mean High Water Level (*MHWL*) were also assessed. The *MHWL* is defined as  
350 the average of all the water level maxima observed during the study period and it thus represents  
351 a meaningful proxy to estimate changes in flooding risk in urban areas within the lagoon.  
352 Overall, a generalized increase in *MHWL* occurred during the study period (Figure 5a-f). A  
353 slight attenuation of *MHWL* is observed between 1901 and 1932 (Figure 5k), followed by a  
354 pronounced increase between 1932 and 1970 (Figure 5k), when more than half of the total marsh  
355 area was already lost (Figure 2k) and the lagoon underwent significant morphological changes  
356 (Figure 2c,d). After 1970, only minor increases in *MHWL* are observed until 2014 (Figure 5k).  
357 Nonetheless, simulations in the hypothetical scenarios suggest that additional loss of salt marshes  
358 could have resulted in further *MHWL* increases relative to the values observed in 2014 (Figure  
359 5).

### 360 **4.2 Wind waves and bottom shear stresses**

361 Wind waves play a fundamental role in the hydrodynamics and morphodynamics of  
362 shallow tidal systems, in general (e.g., Green and Coco 2014), and of the Venice Lagoon, in  
363 particular (Carniello et al., 2009, 2011, 2012). Here we focus on change in the maximum  
364 significant wave height ( $H_{sMAX}$ , Figure 6).  $H_{sMAX}$  invariably increases through time in all the  
365 considered historical configurations, but only minor changes occurred before 1932 (Figure  
366 6a,b,c,k). In contrast, between 1932 and 1970, pronounced increases in  $H_{sMAX}$  are observed,  
367 especially in the central and southern portions of the lagoon that are more exposed to the action  
368 of *Bora* winds (Figure 6d). Although after 1970 the distribution of  $H_{sMAX}$  does not display

369 substantial changes (Figure 6e,f,k), numerical simulations considering additional loss of salt  
 370 marshes suggest that  $H_{sMAX}$  would have further increased proportionally to the percentage of  
 371 marsh area being lost (Figure 6g-j and l). The most important increases in  $H_{sMAX}$  would have  
 372 occurred in areas that are presently occupied by extensive salt marshes (Figure 6g-j), that is, the  
 373 whole northern lagoon as well as the most landward portions of the central-southern lagoon (see  
 374 Figure 2f).

375 The key role exerted by wind waves on the lagoon morphodynamics is related to their  
 376 ability to determine sediment resuspension from shallow tidal-flat areas, a process whose  
 377 intensity depends nonlinearly on the wave characteristics. Wind waves produce bottom shear  
 378 stresses ( $\tau_{ww}$ ) that compound the bottom shear stresses induced by tidal currents ( $\tau_b$ ) and  
 379 determine the total shear stresses ( $\tau_{wc}$ ) that eventually lead to sediment resuspension when  $\tau_{wc}$   
 380 exceeds the critical threshold for erosion (Carniello et al. 2012; see section 3.1). Numerical  
 381 results suggest that within tidal channels, where  $\tau_b$  is typically dominant, only minor increases in  
 382  $\tau_{wc}$  maxima occurred over time (Figure 7a-f). In contrast, across shallow mudflats where the  
 383 wave-induced bottom shear stress component ( $\tau_{ww}$ ) is predominant, the maximum values of  $\tau_{wc}$   
 384 invariably increases from 1887 to 2014 (Figure 7a-f). This is consistent with changes in  
 385 maximum wave heights ( $H_{sMAX}$ ), and results into a generalized increase of  $\tau_{wc}$  across the entire  
 386 lagoon that is especially clear between 1932 and 1970 (Figure 7k). Numerical simulations with  
 387 additional losses of salt marshes suggest that  $\tau_{wc}$  maxima would have been further enhanced  
 388 proportionally to the marsh area being lost (Figure 7g-j,l).

### 389 4.3 Tidal asymmetries

390 Larger values of bottom shear stress increase the chance for sediments to be resuspended  
 391 and transported elsewhere by tides and wave-induced currents, thus affecting the overall lagoon  
 392 sediment budget. In particular, previous studies showed that asymmetries in tidal currents  
 393 (Aubrey & Speer, 1985; Murty, 1990) are critical in determining the ultimate fate of sediments  
 394 carried in suspension, with ebb-dominated tidal flow leading to sediment export to the open sea  
 395 and, therefore, to a net erosion of the lagoon. Here we quantified tidal asymmetries ( $\gamma$ ) using the  
 396 formulation proposed by Nidzieko (2010). This formulation allows for a spatially-explicit  
 397 computation of  $\gamma$  based on the normalized skewness of the water level time derivative ( $\partial\zeta/\partial t =$   
 398  $\zeta'$ ):

$$399 \quad \gamma = \frac{\mu_3}{\sigma^3} = \frac{\frac{1}{T-1} \sum_{t=1}^T (\zeta'_t - \bar{\zeta}')^3}{\left[ \frac{1}{T-1} \sum_{t=1}^T (\zeta'_t - \bar{\zeta}')^2 \right]^{3/2}}$$

400 where  $\mu_3$  is the third sample moment about the mean,  $\sigma$  is the standard deviation, and  $T$  is the  
 401 sampling timeframe. Negative values of  $\gamma$  indicate ebb dominance, whereas flood-dominated  
 402 tides are characterized by positive values of  $\gamma$ . Our results suggest that progressively larger  
 403 portions of the Venice Lagoon became ebb-dominated over time (Figure 8a-f). Pronounced  
 404 changes in tidal asymmetry in the surroundings of the Lido inlet can be observed between 1887  
 405 and 1901 (Figure 8b), immediately after the construction of the jetties. Similarly, an extensive  
 406 expansion of the areas dominated by ebb tides is highlighted after the construction of the jetties  
 407 at the Chioggia inlet in 1932 (Figure 8c). Afterward, the hydrodynamic regime of many portions  
 408 of the lagoon shifted from flood- to ebb-dominated, especially close to the Malamocco inlet

409 where the Malamocco-Marghera canal was excavated (Figure 8d-f). The most pronounced  
410 changes in the hydrodynamic regime of the lagoon occurred between 1932 and 1970 (Figure 8k),  
411 when most of the salt marshes had already been lost and the deepening rate of tidal flats  
412 accelerated (see Figure 2k-l). Numerical simulations demonstrate that the effects of additional  
413 marsh losses on  $\gamma$  would have not been negligible (Figure 8g-j). Overall, ebb dominance is  
414 slightly reduced as salt marshes are progressively eroded (Figure 8l), but distinct trends of  
415  $\gamma$  changes are observed as a function of the distance from the lagoon inlets. Specifically, while  
416 ebb dominance is either maintained or enhanced in the portions of the lagoon proximal to the  
417 inlets, a shift to flood dominance is observed in most distal regions where extensive salt marshes  
418 are still found nowadays (Figure 8g-j).

419

## 420 **5 Discussion**

421 Our analyses highlight the difficult task of unraveling the hydrodynamic consequences of  
422 salt-marsh loss in shallow back-barrier systems. Conceptualized scenarios that assume additional  
423 marsh losses without any further modification of the lagoon morphology help to isolate the direct  
424 effects of marsh disappearance on the lagoon hydrodynamics. In contrast, indirect, cascade  
425 effects due to morphodynamics feedbacks are more easily understood based on numerical results  
426 obtained for the historical configurations of the Venice Lagoon.

### 427 **5.1 Water Levels**

428 How marsh disappearance affects water levels within the lagoon is perhaps the most  
429 controversial point to debate. This is because salt marshes are most effective in regulating water  
430 levels in the upper intertidal frame due to their characteristic topographic elevations that exceed  
431 the mean sea level. Conversely, at lower water stages, marshes are not typically flooded and thus  
432 have limited effects on tide propagation, which takes place predominantly within major tidal  
433 channels and across tidal flats.

434 Our simulations of hypothetical marsh-loss scenarios (E25 to E100 runs, Figure 4l) suggest that  
435 marsh erosion overall increases accommodation in the back-barrier system, thus reducing the  
436 average amplitude of tidal oscillations and, therefore, the mean tidal range (*MTR*). This is in  
437 agreement with the results of the numerical experiments carried out in other back-barrier tidal  
438 systems along the continental US coast (e.g., Donatelli et al. 2018). However, our analyses also  
439 highlight a continued increase in *MTR* (Figure 4k) during the last century. This is most likely a  
440 result of both anthropogenic modifications of the lagoon inlets and reduced bottom friction due  
441 to mudflat deepening (e.g., D'Alpaos and Martini 2005; Tambroni and Seminara 2006; Carniello  
442 et al. 2009; Ferrarin et al. 2015) (Figure 2a-f). Therefore, enhanced *MTR* is only partially related,  
443 both directly and indirectly, to salt-marsh loss. This is supported by the local increase of *MTR*  
444 after the construction of jetties at the inlets of Lido (1901, Figure 4b) and Chioggia (1932, Figure  
445 4c). Besides, generalized increases in *MTR* occurred between 1932 and 1970 (Figure 4d) as a  
446 consequence of tidal-flat deepening. Excavation of the Vittorio Emanuele and Malamocco-  
447 Marghera waterways also likely contributed to enhance *MTR*, especially in the central part of the  
448 lagoon (Figure 4c,d). Finally, slight reductions in *MTR* between 2003 and 2014 are related to  
449 enhanced hydraulic resistance at the inlets produced by the Mo.S.E. structures (Figure 4k)  
450 (Maticchio et al., 2017), and are therefore not directly linked to changes in salt-marsh extent.

451 Similarly, the continued increase in mean high water levels (*MHWL*) observed since  
 452 1887 should be only partially considered as a direct effect of marsh loss (Figure 5a-f). Our  
 453 simulations suggest that progressive, additional loss of salt marshes would have resulted in  
 454 further *MHWL* increases (Figure 5g-j). This is most probably related to the progressive fading of  
 455 energy dissipations produced by the presence of marshes at high-water stages. Reduced energy  
 456 dissipations for high-water levels magnify tidal peaks, thus enhancing *MHWL*. Larger *MHWL*  
 457 potentially bears negative consequences in terms of increased flooding risk of urban settlements  
 458 (Ferrighi, 2005; Gambolati & Teatini, 2014; Rinaldo et al., 2008). However, one should  
 459 appreciate that marsh loss preferentially leads to larger *MHWL* in the innermost portions of the  
 460 northern and central-southern lagoon where marshes are still widespread nowadays (Figure 5g-j).  
 461 Conversely, limited changes in *MHWL* are observed in the proximity of the inlets and around the  
 462 major urban settlements, namely Venice city, Chioggia, and the islands of Murano, Burano, and  
 463 Sant'Erasmus. For these locations, changes in *MHWL* are less than 3% even if marshes would  
 464 entirely disappear (E100 scenario), with even lower variations (<2%) in terms of maximum high-  
 465 water levels. Hence, the effects of salt-marsh loss on back-barrier hydrodynamics appear to be  
 466 extremely site-specific, with significant hydrodynamic modifications being observed over  
 467 distances of a few kilometers even within a given tidal embayment.

468

## 469 5.2 Wind waves and bottom shear stresses

470 The direct effects of marsh erosion on significant wave heights ( $H_{S_{MAX}}$ , Figure 6) and the  
 471 associated bottom shear stresses ( $\tau_{wc}$ , Figure 7) are quite straightforward. Before 1932 the  
 472 generation and propagation of large wind waves were hampered by widespread salt marshes  
 473 (Figure 2a-c) that limited wind fetches, as well as by the reduced mudflat depths which promoted  
 474 significant wave-energy dissipation (Figure 6a-c). From 1970 onwards, in contrast, salt-marsh  
 475 losses coupled with tidal-flat deepening led to increasing  $H_{S_{MAX}}$  (Figure 6d-f) (Fagherazzi et al.  
 476 2006; Defina et al. 2007). Reduced marsh coverage favored larger  $H_{S_{MAX}}$  due to longer fetches,  
 477 as confirmed also by numerical simulations involving additional marsh losses (Figure 6g-j). The  
 478 most significant increases in  $H_{S_{MAX}}$  would have occurred in areas that are nowadays occupied by  
 479 extensive salt marshes, that is, the northern lagoon as well as the innermost portions of the  
 480 central-southern lagoon (see Figure 2f). Here, the disappearance of salt marshes would have  
 481 reduced their wind-sheltering effect and lengthened wind fetches, eventually leading to the  
 482 generation and propagation of higher waves (Figure 6g-j).  
 483 Larger  $H_{S_{MAX}}$  also have negative implications for flooding risk in topographically-depressed  
 484 urban areas, which may be exacerbated when bank overtopping occurs (Gambolati & Teatini,  
 485 2014; Mel, Carniello, et al., 2021; Mel, Viero, et al., 2021; Ruol et al., 2020). Nevertheless, our  
 486 analyses show that  $H_{S_{MAX}}$  and the associated flooding risk are not likely to increase in the marsh-  
 487 erosion scenarios near the major urban settlements. Even assuming the total disappearance of salt  
 488 marshes (E100 scenario), increments lower than 5% in  $H_{S_{MAX}}$  are observed near Venice City and  
 489 the inhabited centers of Chioggia, Murano, Burano, and Sant'Erasmus. This is because all these  
 490 locations lie far from extensive salt-marsh areas and are little affected by the salt-marsh losses. In  
 491 contrast, at the landward margins of the lagoon where extensive salt-marsh areas are still found  
 492 nowadays, additional marsh erosion could have led to a threefold to fourfold increase in  $H_{S_{MAX}}$ .  
 493 Values of  $H_{S_{MAX}} > 60$  cm compared to  $H_{S_{MAX}} = 15-20$  cm in the pre-erosion condition confirm

494 the ability of salt marshes to dissipate wave energy and reducing flooding risk in the lagoon most  
495 marginal areas.

496 Notably, higher waves are also likely to threaten the conservation of the remaining salt-marsh  
497 ecosystems, due to the positive feedback mechanism between marsh lateral erosion and wind-  
498 wave power (Carniello et al., 2016; A. D'Alpaos et al., 2013; Finotello et al., 2020; Leonardi,  
499 Defne, et al., 2016; Marani, D'Alpaos, et al., 2011; Tommasini et al., 2019). Specifically, salt-  
500 marsh lateral retreat rates are linearly correlated to wave power ( $P_w$ ) (Finotello et al., 2020;  
501 Leonardi, Ganju, et al., 2016; Marani, D'Alpaos, et al., 2011; Tommasini et al., 2019), which in  
502 turn is a quadratic function of wave height. Hence, salt-marsh loss leads to higher, more  
503 energetic waves, which in turn enhance marsh lateral retreat even further in a superlinear fashion.  
504 In addition, higher waves also produce larger bottom shear stresses, especially across extensive  
505 tidal-flat areas (Figure 7). Our simulations demonstrate that additional losses of salt marshes  
506 would have further enhanced  $\tau_{wc}$  proportionally to the marsh area being lost (Figure 7g-j,l). This  
507 is because the wind sheltering effect typically offered by marshes would have been progressively  
508 reduced, allowing for increasingly higher waves to winnow the lagoon bottom (Carniello et al.,  
509 2014, 2016; A. D'Alpaos et al., 2013; Tommasini et al., 2019). From a morphodynamic  
510 standpoint, the implications of increasing  $\tau_{wc}$  can be manifold. Larger  $\tau_{wc}$  will enhance sediment  
511 entrainment from the lagoon shallows, producing higher concentrations of suspended sediment  
512 (*SSC*) (Tognin et al., 2022). Wave-driven resuspension from tidal flats represents a key source of  
513 sediment for salt marshes in sediment-starving shallow tidal embayments, where the majority of  
514 mineral sediments are delivered to the marsh surface during storm-surge events concomitant with  
515 strong wave activity (Tognin et al., 2021). Thus, enhanced *SSC* could ensure higher resilience of  
516 salt-marsh ecosystems in the face of rising relative sea levels (Elsley-Quirk et al., 2019; Mariotti  
517 & Fagherazzi, 2010; Tognin et al., 2021). However, such a beneficial effect is likely offset by  
518 marsh loss via lateral retreat, which would reduce the total marsh area and promote  
519 fragmentation, in this way hampering the marsh's ability to capture suspended sediment and cope  
520 with sea-level rise (Donatelli, Zhang, et al., 2020; Duran Vinent et al., 2021). Besides, enhanced  
521 *SSC*, coupled with the generally ebb-dominated character of tides (Figure 8 and see Section 5.3),  
522 are likely to negatively affect the lagoon net-sediment budget, leading to further tidal flat  
523 deepening and salt marsh losses.

524

### 525 **5.3 Tidal asymmetries**

526 Salt-marsh loss affects tidal asymmetry ( $\gamma$ ) mostly in an indirect fashion, with manmade  
527 modifications on the lagoon inlets playing in contrast a critical role (L. D'Alpaos, 2010; L.  
528 D'Alpaos & Martini, 2005; Matticchio et al., 2017; Tambroni & Seminara, 2006). Indeed, ebb  
529 dominance was increased after the completion of the jetties both at the Lido (Figure 8b) and  
530 Chioggia inlets (Figure 8c). Indirect effects of salt-marsh loss on  $\gamma$  could instead arise from the  
531 positive morphodynamic feedback between marsh loss and tidal-flat deepening (Carniello et al.,  
532 2008, 2009; Defina et al., 2007). This feedback is likely responsible for the shift from flood- to  
533 ebb-dominance observed in many portions of the lagoon between 1932 and 1970, especially in  
534 the area facing the Malamocco inlet where the Malamocco-Marghera shipway was excavated  
535 (Ferrarin et al., 2015) (Figure 8d-e). This speculation is supported by earlier results showing how  
536 deeper tidal flats reduce bottom friction and ultimately enhance ebb-dominance (Dronkers, 1986;  
537 Fortunato & Oliveira, 2005; Friedrichs, 2012; Guo et al., 2019). Our numerical simulations

538 demonstrate also that the direct effects of additional marsh losses on  $\gamma$  are potentially not  
539 negligible, with progressive marsh erosion leading to more widespread flood-dominated areas in  
540 the innermost portions of the lagoon (Figure 8g-j). This result is consistent with evidence from  
541 previous studies showing that decreasing intertidal storage capacity associated with gradual  
542 marsh losses would re-establish the flood dominance typically observed for progressive tidal  
543 waves (Dronkers, 1986; Rinaldo et al., 1999). Despite increased flood dominance due to the  
544 marsh disappearance, the tidal regime in most of the lagoon would have remained ebb-dominated  
545 (Figure 8g-j). Thus, a reversal of the current erosional trend and the establishment of a positive  
546 net sediment budget would have been unlikely. However, given the non-linear dependence of  
547 sediment transport processes on both tidal flow velocities and asymmetry, these hypotheses  
548 should be verified by *ad hoc* morphodynamic simulations.

549

#### 550 **5.4 Implications for the hydrodynamics of back-barrier tidal lagoons**

551 Our analyses highlight both direct and indirect effects of salt-marsh deterioration on the  
552 hydrodynamics of the Venice Lagoon. Despite being generally consistent with previous studies  
553 carried out in different tidal settings (e.g., Donatelli et al. 2018, 2020b, a), care should be given  
554 to generalizing our results to back-barrier tidal embayments morphologically and  
555 hydrodynamically different from the study case at hand. The reasons behind this caution are  
556 manifold, though all broadly related to the morphological and hydrodynamic peculiarities that  
557 characterize each tidal environment, as well as to the conceptualizations we adopted in our  
558 numerical simulations.

559 First, the hydrodynamic response of a tidal system to marsh erosion depends on i) the  
560 planform geometry and hypsometry of the basin (Deb et al., 2022; Van Maanen et al., 2013); ii)  
561 the characteristics of tides, especially in terms of tidal range and progressive vs. standing  
562 character of the system (Van Maanen et al., 2013; Ward et al., 2018; Zhou et al., 2018); and iii)  
563 the wave climate, affecting both the basin hydrodynamics and its sediment transport regime  
564 (Carniello et al., 2011; A. D'Alpaos et al., 2013). Particularly important is the spatial distribution  
565 of salt marshes within the back-barrier basin. As demonstrated by Donatelli et al. (2020b),  
566 different hydrodynamic changes due to marsh loss are to be expected in back-barrier systems  
567 where most marshes fringe the mainland compared to those characterized by the presence of  
568 extensive marsh areas detached from the mainland.

569 Second, our simulated scenarios do not account for salt-marsh landward migration, which  
570 can mitigate the net loss of salt marshes (e.g., Feagin et al. 2010; Field et al. 2016; Enwright et  
571 al. 2016; Fagherazzi et al. 2019; Kirwan and Gedan 2019). Even though this process is typically  
572 hindered by the presence of levees and dikes at the interface between marshes and the upland  
573 (e.g., Yang et al., 2022), it cannot be disregarded *a priori*. The colonization of new intertidal  
574 areas by marsh upland migration would profoundly change the hydrodynamics and sediment  
575 budget of the whole back-barrier system, as new areas would be periodically flooded by tides  
576 and additional sediment volumes would become available as marshes expand landward.

577 Finally, the timescale required for the system to morphodynamically adapt to changes in  
578 marsh coverage is generally difficult to quantify. This is because sediment volumes made  
579 available by marsh lateral erosion can be redistributed within the basin, thus affecting both the  
580 net sediment budget and the related morphological changes (Donatelli, Kalra, et al., 2020; Elsey-

581 Quirk et al., 2019; Kalra et al., 2021). In our simulations, sediments eroded from marshes are  
582 instantaneously removed and can no longer contribute to the lagoon's morphological evolution.  
583 Moreover, it should be noted that salt-marsh loss seldom occurs without inducing modifications  
584 to other back-barrier landforms. This is clearly highlighted by historical field data from the  
585 Venice Lagoon, which suggests mutual feedback between marsh erosion and tidal-flat  
586 deepening. Therefore, generalizations of results obtained from hypothetical erosive scenarios  
587 should be treated with caution, since the modeled hydrodynamic changes could be mitigated or  
588 magnified by other morphodynamic adjustments induced by marsh disappearance on the tidal  
589 back-barrier system as a whole.

590 In view of the above, the effects of salt-marsh loss on the hydrodynamic and  
591 morphodynamics of shallow back-barrier tidal systems are likely to be extremely site-specific,  
592 and therefore difficult to generalize. Moreover, biogeomorphological feedbacks, which are key  
593 drivers of marsh spatiotemporal evolution, are likely to vary geographically as a consequence of  
594 distinct ecological community assemblages and different climatic forcings. This further  
595 complicates predicting the morphodynamic effects of salt-marsh loss, as well as the timescales  
596 over which such effects would manifest (Bertness & Ewanchuk, 2002; Finotello, D'Alpaos, et  
597 al., 2022; Pennings & He, 2021; Wilson et al., 2022).

## 598 **6 Conclusions**

599 In this study, we focused on the microtidal Venice Lagoon (Italy) to disentangle the role  
600 played by the loss of salt marshes on the hydrodynamics of tidal back-barrier embayments.  
601 Numerical simulations were performed considering both past morphological configurations of  
602 the lagoon dating back up to 1887 and hypothetical scenarios involving additional marsh erosion  
603 relative to the present-day conditions. This allowed us to highlight both the direct and indirect  
604 effects of salt-marsh loss on the evolution of the lagoon hydrodynamics. Direct effects include  
605 enhanced mean-high water levels due to reduced energy dissipation at high-water stages, and the  
606 formation of higher and more powerful wind waves due to longer fetches. Moreover, historical  
607 data and numerical results suggest that marsh disappearance is likely to trigger tidal-flat  
608 deepening, thus leading to increased tidal ranges due to reduced energy dissipation and  
609 modifications of local tidal asymmetries. We also speculated on the potential impacts of the  
610 observed hydrodynamic changes on the lagoon ecomorphodynamic evolution, as well as on the  
611 associated risk of tidal flooding in urban settlements. Our analyses suggest that a failure to  
612 restore and protect the existing marshes in the past would have not critically affected the  
613 flooding risk in urban lagoonal settlements. On the contrary, flooding risk would have been  
614 significantly increased in the lagoon marginal areas, mostly due to larger wave heights.  
615 Our findings provide novel insights into the hydrodynamic effects of salt-marsh loss in sediment-  
616 starving, shallow tidal embayments morphodynamically dominated by wind-driven sediment  
617 transport processes, with far-reaching implications for the conservation and restoration of coastal  
618 ecosystems that extend well beyond the study case at hand. However, we stress that care should  
619 be given to generalizing the results presented here to tidal embayments that are morphologically  
620 and morphodynamically different from the Venice Lagoon. We support the idea that the response  
621 of back-barrier systems to changing external forcing is highly dependent on site-specific  
622 morphological and ecological features, as well as on the site-specific characteristics of tides,  
623 waves, and sediments.

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630 **Author contributions**

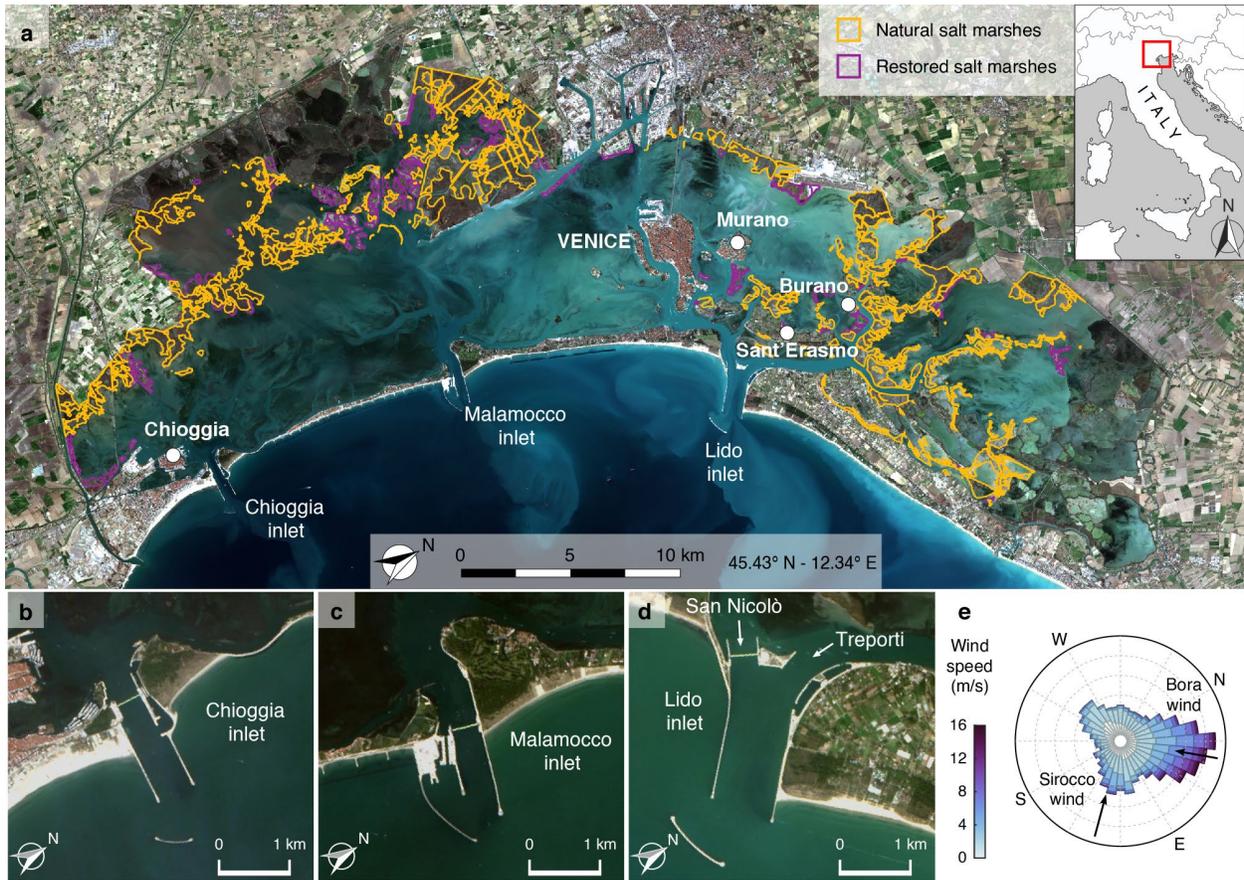
631 Conceptualization: Alvise Finotello, Davide Tognin, Andrea D'Alpaos, Luca Carniello;  
632 Methodology: Alvise Finotello, Davide Tognin, Andrea D'Alpaos, Luca Carniello;  
633 Formal analysis and investigation: Alvise Finotello, Davide Tognin;  
634 Figures: Davide Tognin;  
635 Writing - original draft preparation: Alvise Finotello, Davide Tognin;  
636 Writing - review and editing: all authors;  
637 Funding acquisition: Andrea D'Alpaos, Luca Carniello, Enrico Bertuzzo;  
638 Resources: Andrea D'Alpaos, Luca Carniello, Enrico Bertuzzo, Massimiliano Ghinassi;  
639 Supervision: Andrea D'Alpaos, Luca Carniello, Enrico Bertuzzo.

640 **Open Research**

641 All data needed to evaluate the results presented in the paper can be found at  
642 <http://researchdata.cab.unipd.it/id/eprint/646>. Meteorological data for the Venice Lagoon are also  
643 freely available at [www.comune.venezia.it/content/dati-dalle-stazioni-rilevamento](http://www.comune.venezia.it/content/dati-dalle-stazioni-rilevamento) and  
644 [www.venezia.isprambiente.it/rete-meteo-mareografica](http://www.venezia.isprambiente.it/rete-meteo-mareografica). Data regarding the morphology of the  
645 Venice Lagoon, including those related to salt-marsh coverage, can be found at  
646 <http://www.atlantedellalaguna.it/>

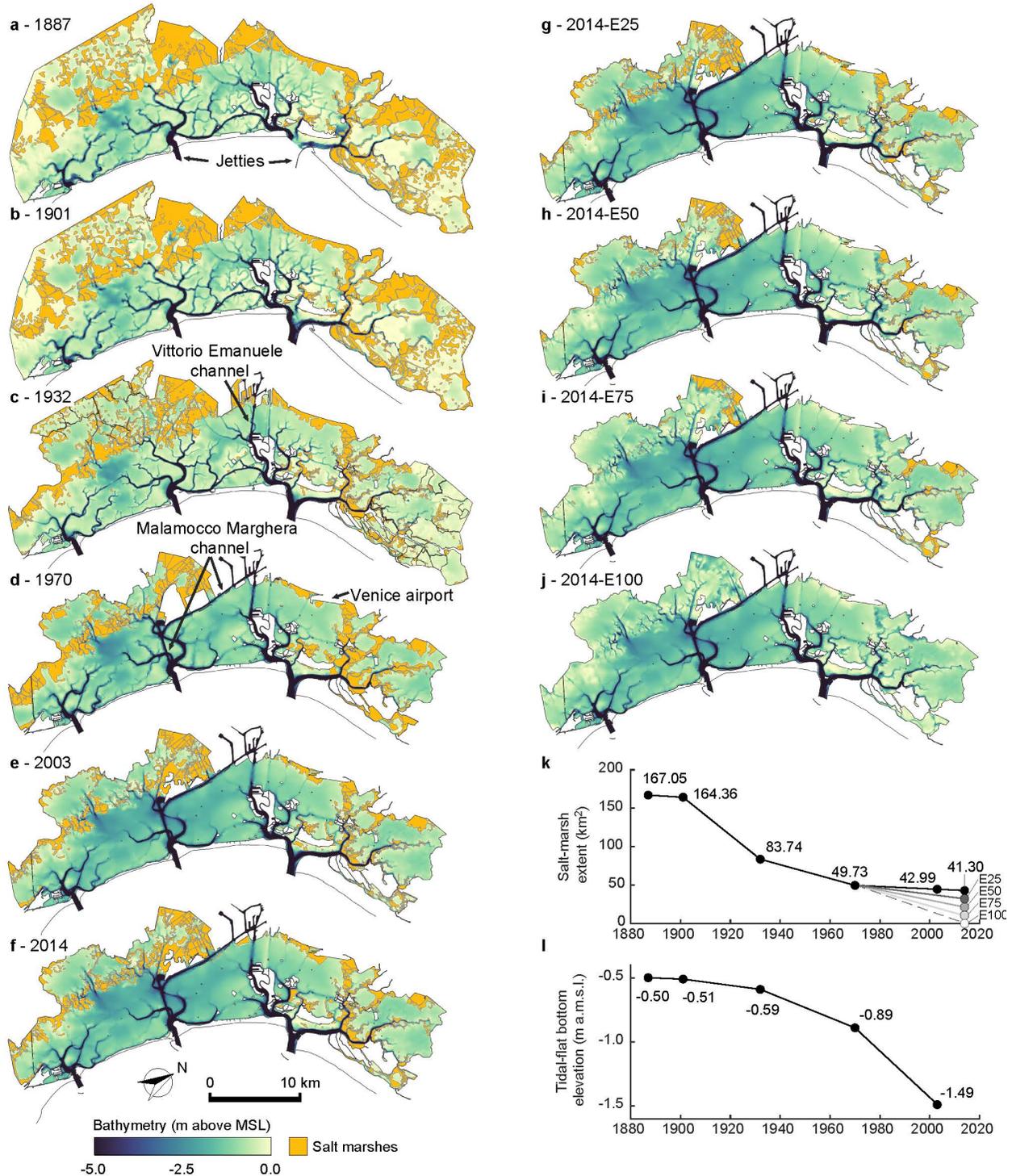
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648 **Figures**



649

650 **Figure 1:** Geomorphological setting. (a) Satellite images of the Venice Lagoon (image Copernicus Sentinel, 2020).  
 651 Natural salt marshes are bordered in yellow, while restored salt marshes are shown in purple. (b, c, d) Close-up  
 652 views of the three lagoon inlets. (e) Rose-diagram representation of wind climate recorded at the “Chioggia Diga  
 653 Sud” anemometric station during the period 2000-2019. The two most relevant winds, i.e., the north-easterly *Bora*  
 654 wind and south-easterly *Sirocco* wind, are also highlighted.



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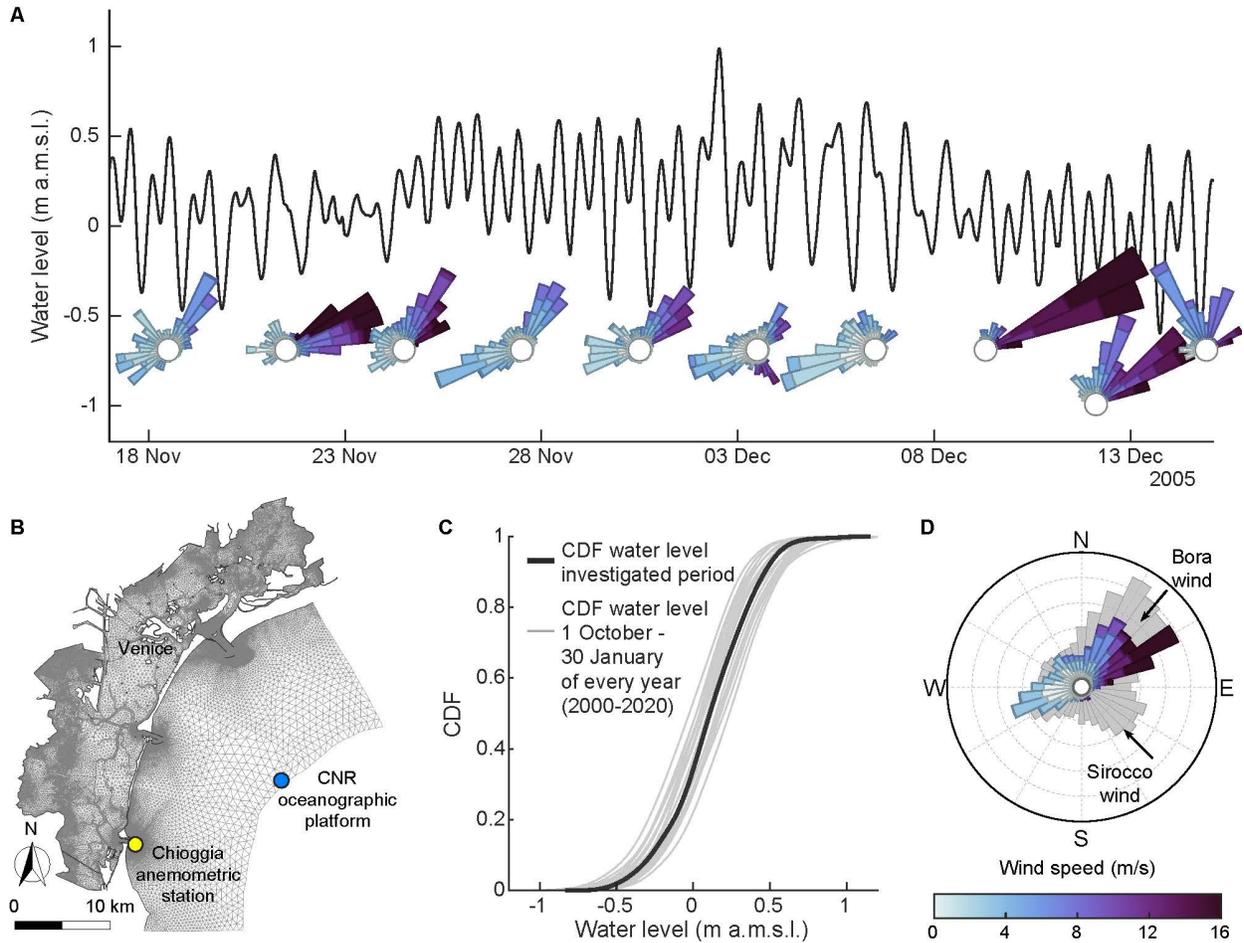
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**Figure 2:** Morphological evolution of the Venice Lagoon. Bathymetry of the Venice Lagoon in 1887 (a), 1901 (b), 1932 (c), 1970 (d), 2003 (e), and 2014 (f), as reconstructed from available historical topographic and bathymetric data. (g,h,i,j) Morphology of the Venice Lagoon according to the hypothetical scenarios of marsh erosion analyzed in the present study. These scenarios assume different rates of marsh erosion equal to 25% (2014-E25), 50% (2014-E50), 75% (2014-E75), and 100% (2014-E100), respectively. Temporal variation of salt-marsh extent (k) and spatially-averaged bottom elevation of tidal flats (l) between 1887 and 2014 are also shown.



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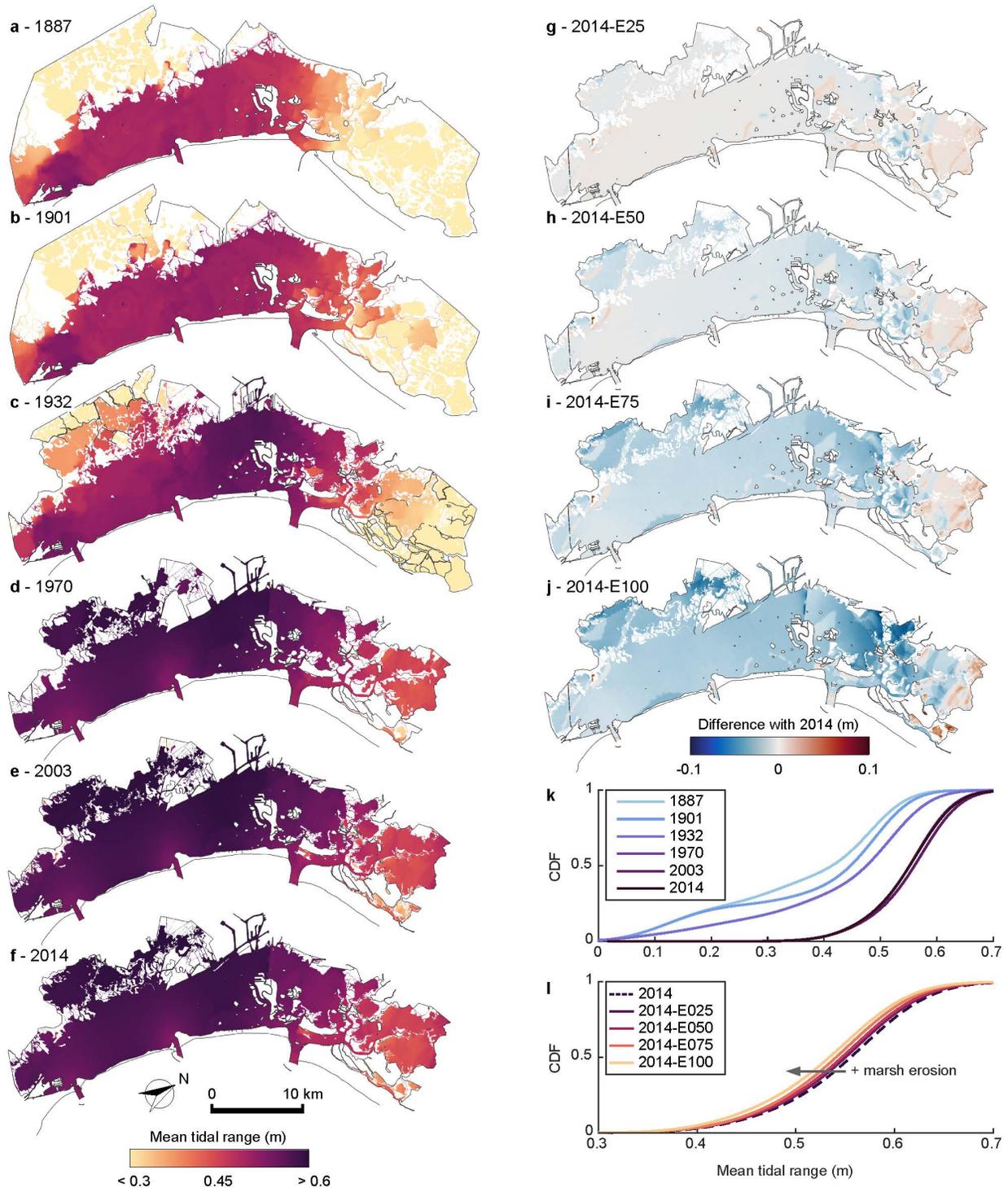
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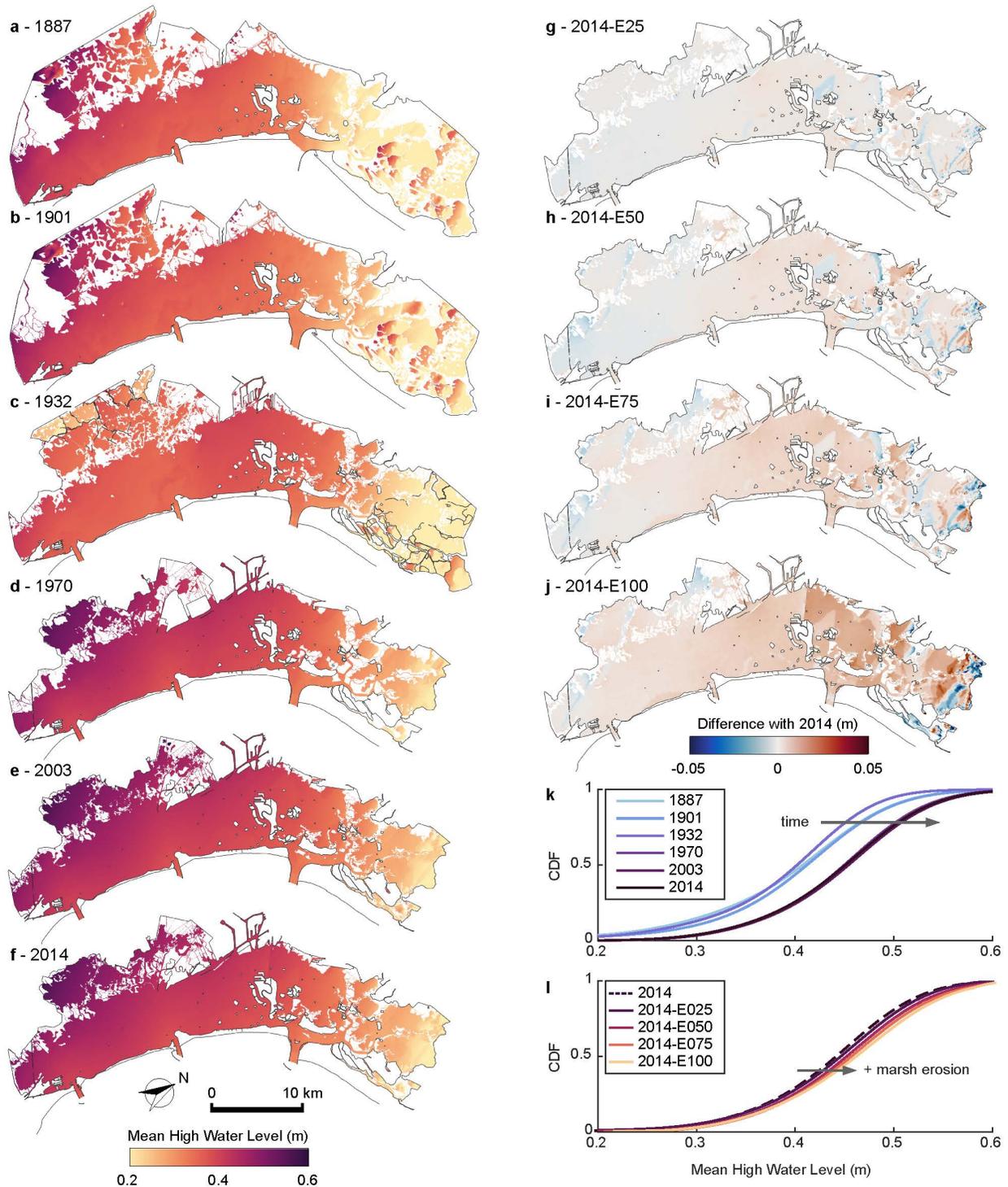
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**Figure 3:** Numerical modeling and boundary conditions. (a) Water level and wind climate data utilized in the numerical simulations. Data refer to the period 17 November 2005 – 17 December 2005. Water levels were measured at the “CNR Oceanographic Platform”, whereas wind data were retrieved from the “Chioggia Diga Sud” anemometric station (see panel b). (b) An example of the computational grid employed by the numerical model, referred to the 2014 morphological configuration of the Venice Lagoon. (c,d) Distributions of water levels (c) and wind climate (d) during the analyzed period are compared to those observed over the period 2000-2019 (in grey).



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**Figure 4:** Evolution of mean tidal range (*MTR*). Spatially-explicit representation of *MTR* in 1887 (a), 1901 (b), 1932 (c), 1970 (d), 2003 (e), 2014 (f). Difference between the 2014 configuration and the hypothetical marsh-erosion scenarios: 2014-E25 (g), 2014-E50 (h), 2014-E75 (i), 2014-E100 (j). (k) Cumulative frequency (CDF) of *MTR* for the historical configurations (1887-2014). (l) Cumulative frequency of *MTR* in the hypothetical marsh erosion scenarios (2014-E25, E50, E75, E100).



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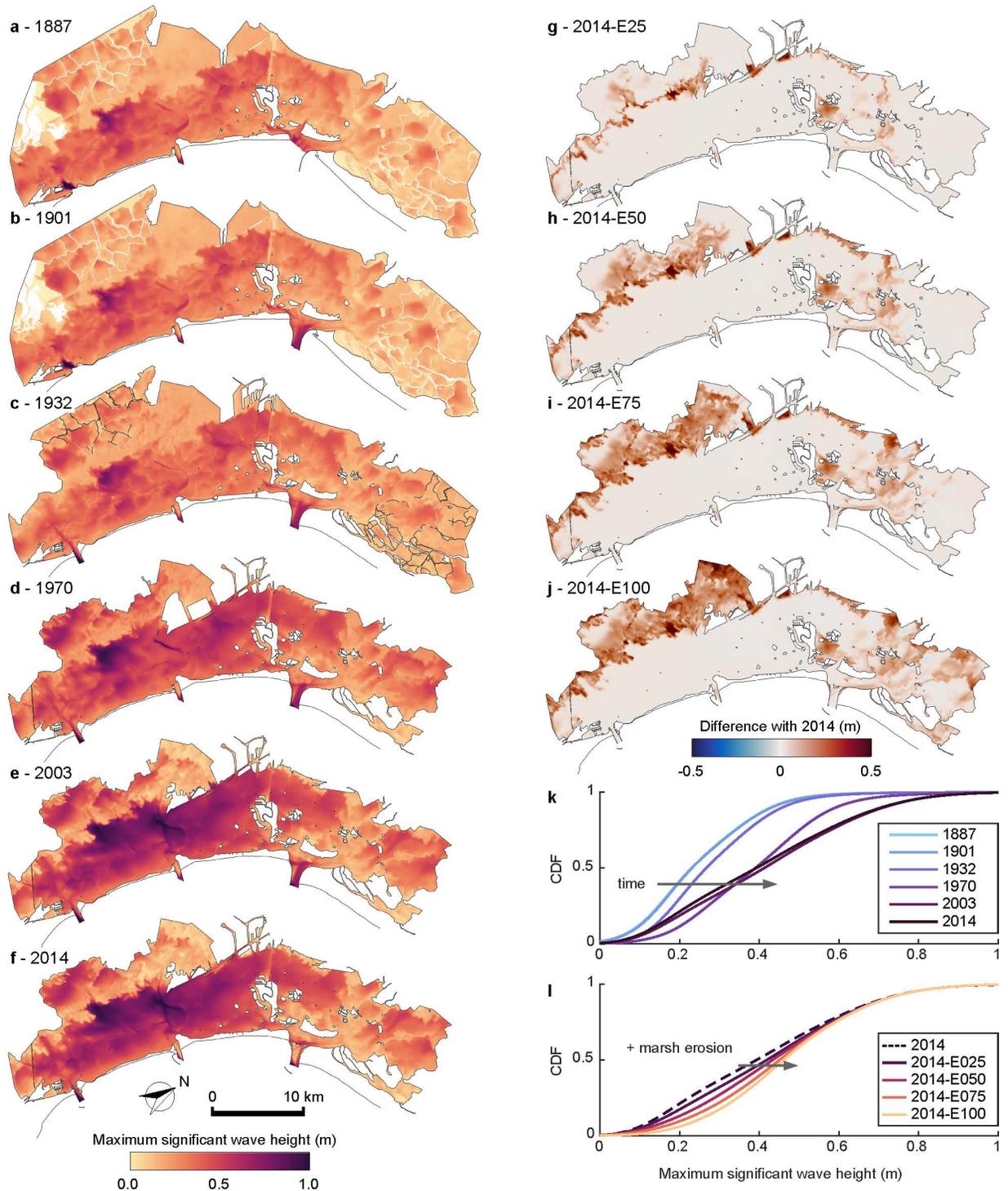
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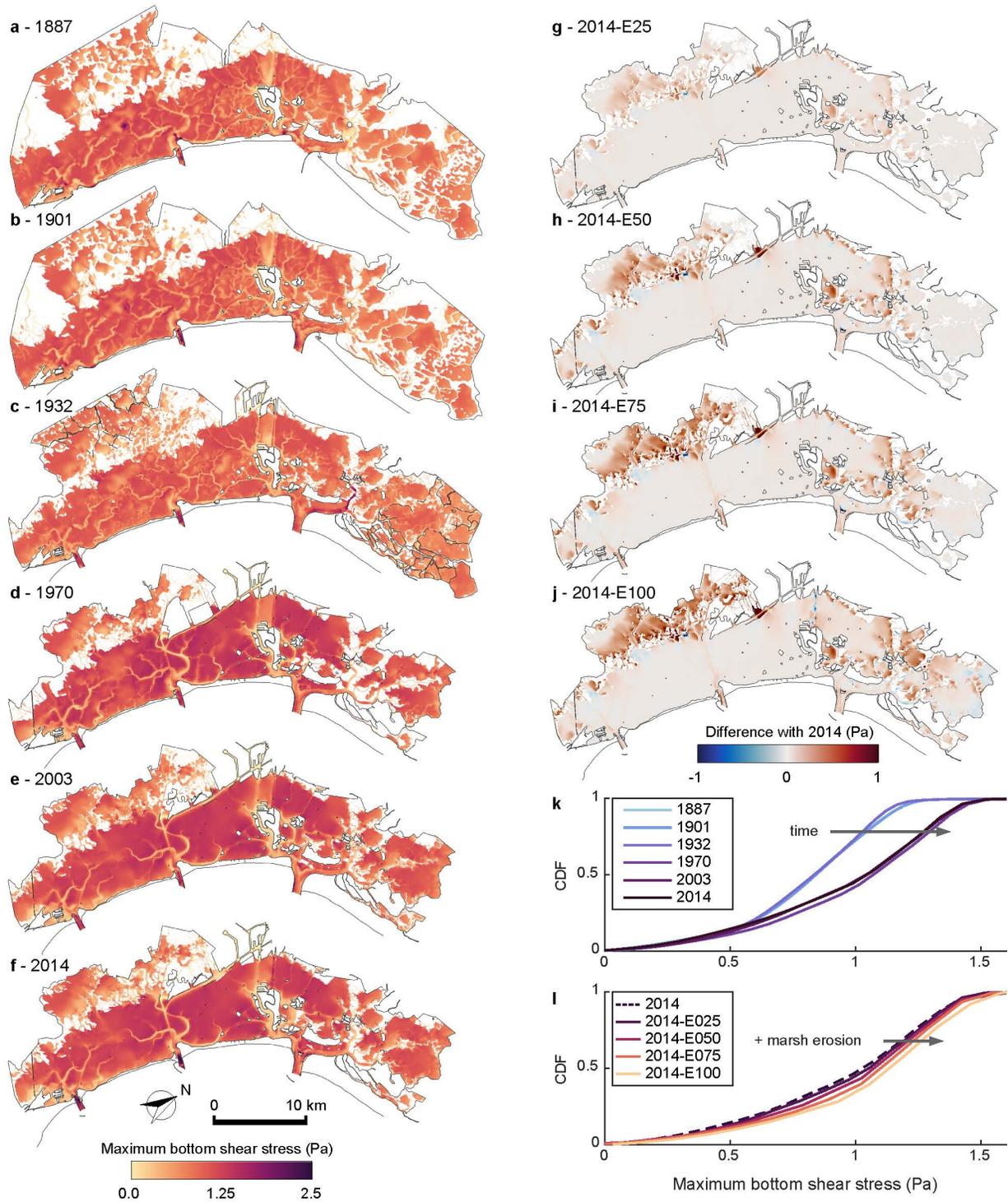
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**Figure 5:** Evolution of mean-high water level (*MHWL*). Spatially-explicit representation of *MHWL* in 1887 (a), 1901 (b), 1932 (c), 1970 (d), 2003 (e), 2014 (f). Difference between the 2014 configuration and the hypothetical marsh-erosion scenarios: 2014-E25 (g), 2014-E50 (h), 2014-E75 (i), 2014-E100 (j). (k) Cumulative frequency (CDF) of *MHWL* for the historical configurations (1887-2014). (l) Cumulative frequency of *MHWL* in the hypothetical marsh erosion scenarios (2014-E25, E50, E75, E100)



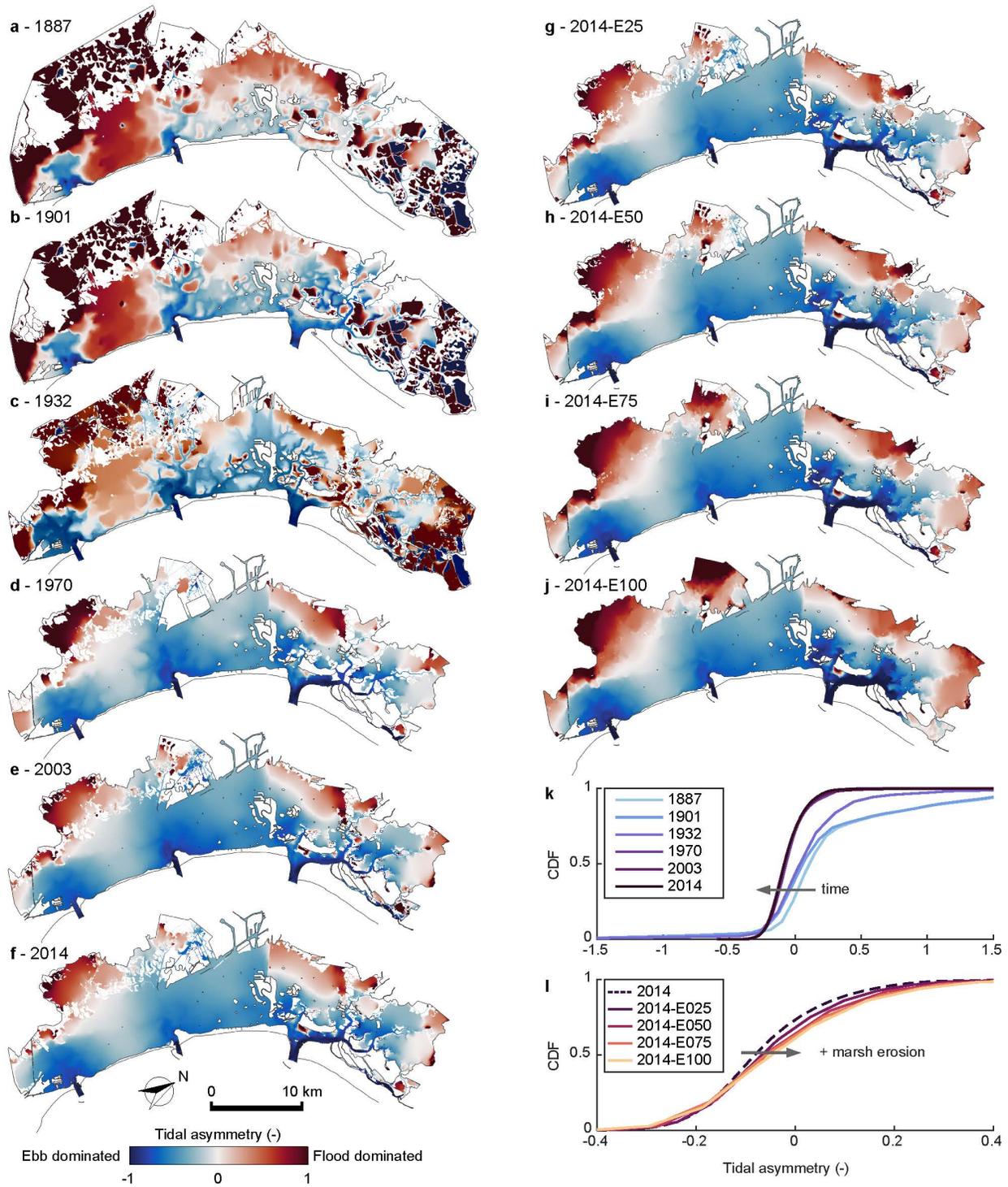
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683 **Figure 6** Evolution of maximum significant wave height ( $H_{smax}$ ). Spatially-explicit representation of  $H_{smax}$  in 1887  
 684 (a), 1901 (b), 1932 (c), 1970 (d), 2003 (e), 2014 (f). Difference between the 2014 configuration and the hypothetical  
 685 marsh-erosion scenarios: 2014-E25 (g), 2014-E50 (h), 2014-E75 (i), 2014-E100 (j). (k) Cumulative frequency  
 686 (CDF) of  $H_{smax}$  for the historical configurations (1887-2014). (l) Cumulative frequency of  $H_{smax}$  in the hypothetical  
 687 marsh erosion scenarios (2014-E25, E50, E75, E100).



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**Figure 7:** Evolution of maximum bottom shear stress ( $\tau_{wc}$ ). Spatially-explicit representation of  $\tau_{wc}$  in 1887 (a), 1901 (b), 1932 (c), 1970 (d), 2003 (e), 2014 (f). Difference between the 2014 configuration and the hypothetical marsh-erosion scenarios: 2014-E25 (g), 2014-E50 (h), 2014-E75 (i), 2014-E100 (j). (k) Cumulative frequency (CDF) of  $\tau_{wc}$  for the historical configurations (1887-2014). (l) Cumulative frequency of  $\tau_{wc}$  in the hypothetical marsh erosion scenarios (2014-E25, E50, E75, E100).



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**Figure 8:** Evolution of tidal asymmetry ( $\gamma$ ). Spatially-explicit representation of  $\gamma$  in 1887 (a), 1901 (b), 1932 (c), 1970 (d), 2003 (e), 2014 (f), 2014-E25 (g), 2014-E50 (h), 2014-E75 (i), 2014-E100 (j). (k) Cumulative frequency (CDF) of  $\gamma$  for the historical configurations (1887-2014). (l) Cumulative frequency of  $\gamma$  in the hypothetical marsh erosion scenarios (2014-E25, E50, E75, E100).

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