

1 **Critical zone storage control on the water ages in ecohydrological outputs**

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12 **Key Points**

- 13
- 14 • Age since precipitation displays inverse storage effect in stream, but not
  - 15 transpiration and soil evaporation, in a humid northern catchment
  - 16 • Hysteresis between storage and the age of transpired water suggests cross-season
  - 17 carryover, despite weak hydroclimatic seasonality
  - 18 • Downslope water subsidies result in valley bottom having weaker storage-age
  - 19 relationships than seen in freely-draining hillslopes
- 20
- 21

## 22 **Plain Language Summary**

23 Knowing how much time water spends in a landscape (its “age”) helps understanding how  
24 water travels through it. These dynamics inform of the stability of water resources for  
25 ecosystems and societies, and of their vulnerabilities under climate and land use changes.  
26 Water ages may vary depending on how wet or dry a location gets between seasons and  
27 years. We thus need to learn more about the demographics (“how much and how old?”) of  
28 the water used by plants, evaporated from soils, and flowing in streams, but it is often  
29 impossible to monitor the heterogeneity of water pathways within landscapes. Addressing  
30 this challenge, we used a numerical model built upon coupling ecohydrological processes and  
31 that maps landscape locations. We adjusted this model using multiple datasets in a catchment  
32 representative of humid boreal environments where climate and vegetation are rapidly  
33 changing. We found markedly different aging patterns between water escaping the system  
34 through the plants, soils, and stream, depending on water storage status. This changing  
35 duration of water movement also differs between the catchment as a whole and its parts. This  
36 method can be used to better understand the multiple ways in which water moves through  
37 landscapes, in current and future conditions.

## 38 **Abstract**

39 Spatially-explicit knowledge of the origins of water resources for ecosystems and rivers is  
40 challenging when using tracer data alone. We use simulations from a spatially-distributed  
41 model calibrated by extensive ecohydrological datasets in a small, energy-limited catchment,  
42 where hillslope-riparian dynamics are broadly representative of humid boreal headwater  
43 catchments that are experiencing rapid environmental transition. We hypothesize that in  
44 addition to wetness status, landscape heterogeneity modulates the water pathways that sustain  
45 ecosystem function and streamflows. Simulations show that catchment storage inversely  
46 controls streamwater ages year-round, but only during the drier seasons for transpiration and  
47 soil evaporation. The ages of these evaporative outputs depend much less on wetness status in  
48 the oft-saturated riparian soils than on the freely-draining hillslopes that subsidize them. This  
49 work highlights the need to consider local dynamics and time-changing lateral  
50 heterogeneities when interpreting the ages, and thus the vulnerability, of water resources  
51 feeding streams and ecosystems in landscapes.

## 52 **1 Introduction**

53 The age of water as it is routed along different pathways within a landscape, i.e., the transit  
54 time to exit from entry as zero-aged inputs (e.g., precipitation in catchment hydrology, or  
55 groundwater recharge in hydrogeology), has long been recognized as an important metric that  
56 explicitly indexes linkages between stores and fluxes (Bolin & Rodhe, 1973; Hrachowitz et  
57 al., 2016). In particular, the demographics of the water ultimately used by vegetation or  
58 evaporated from the soil directly informs on the “temporal depth” of the resources supporting  
59 life and the biogeochemical cycles in the shallow underground of the critical zone (Sprenger  
60 et al., 2019). From seasonal carryover of precipitation inputs (Allen et al., 2019; Kuppel et  
61 al., 2017) to vegetation responses to interannual variations in water availability (Bales et al.,  
62 2018; Chitra-Tarak et al., 2018; Hahm et al., 2019), knowing the ages of this “green water”  
63 (i.e., not feeding streamflow nor deep recharge; (Falkenmark & Rockström, 2006) would  
64 crucially help in assessing the vulnerability of natural vegetation communities to drought and  
65 climatic stress, and the sustainability of managed systems (e.g. timber and rain-fed  
66 agrosystems).

67 Quantifying the demographics of green water has however received little attention (Soulsby,  
68 Birkel, et al., 2016), due to the difficulty of directly disentangling the components of green

69 water dynamics outside of carefully controlled experiments (Evaristo et al., 2019). Most of  
70 the analytical developments regarding transit time characterization have remained focused on  
71 dating streamwater or groundwater using concentrations in environmental tracers (Bethke &  
72 Johnson, 2008; McGuire & McDonnell, 2015), for which advanced analytical frameworks  
73 resolve the water age distribution of outputs or of water stores, sometimes linking the two  
74 (Benettin et al., 2013; Botter, 2012; Harman, 2015, 2019; Hrachowitz et al., 2013; Jasechko,  
75 2019; Jasechko et al., 2016; Rinaldo et al., 2015). These efforts have highlighted the dynamic  
76 link between storage dynamics and water ages in streamflow as a key signature of  
77 hydrological functioning, yet there is currently no framework to assess how such  
78 relationships would translate for spatially-distributed outputs such as plant transpiration and  
79 soil evaporation. In most cases water ages are interpreted as a mixture of water being  
80 partitioned between shallower and deeper flow paths (e.g., surface runoff, interflow,  
81 groundwater lateral flow) in a lumped description of the spatial domain. This vertical view,  
82 however, remains to be articulated with the lateral heterogeneity in critical zone attributes  
83 (e.g. land cover type, topographic position, subsurface properties, vegetation phenology,  
84 slope exposure) and resulting variability of pathways, mixing and storages. This will help not  
85 only avoiding misinterpretation of apparent water ages dynamics at spatially aggregated  
86 scales (Kirchner, 2016; Soulsby et al., 2015), but, crucially, better capturing heterogeneous  
87 ecohydrological responses to environmental changes observed within real-world landscapes  
88 (Bales et al., 2018).

89 Addressing the crucial issue of water resources vulnerabilities in the critical zone requires  
90 bridging the wealth of information from ecohydrological measurements of water  
91 composition, fluxes and stores with the capacities of process-based models to inform on  
92 internal hydrological states (Maneta et al., 2018; Wilusz et al., 2020; Yang et al., 2018). We  
93 do so using a novel spatially-distributed model where the time-varying water ages in  
94 ecohydrological fluxes are dynamically derived from mixing equations at the grid cell level.  
95 It is applied in a small (3.2 km<sup>2</sup>) energy-limited, humid headwater catchment where the  
96 model has been calibrated and validated using exceptionally diverse and long-term datasets  
97 (Kuppel et al., 2018b, 2018a). Our analysis shows not only that the seasonal changes in  
98 landscape wetness exert a distinctive control on the ages of water transpired by plants,  
99 evaporated from the soil and flowing in the stream, but also that landscape position and  
100 lateral heterogeneity between hydrogeological units play a role in modulating the  
101 availability of younger replenishable water and older water that renews more slowly. An  
102 implication is that landscape heterogeneity is a strong control on the relative importance of  
103 flow pathways that contribute water for ecosystem health and streamflow generation. The  
104 effect of landscape heterogeneity may thus obscure the actual impact of climate variability on  
105 the water sources that sustain ecosystem function. Our findings may help assess the true  
106 vulnerability of ecosystem water supplies to precipitation variability and land cover change,  
107 thus advancing our understanding of hydrologic resilience in regions experiencing  
108 environmental transition.

## 109 **2 Materials and Methods**

### 110 **2.1 Study catchment**

111 Bruntland Burn is a 3.2 km<sup>2</sup> headwater catchment in the Scottish Highlands (Fig. S1).  
112 Elevation ranges between 220 and 560 m above sea level, with a wide valley bottom and  
113 steep slopes, typical of post-glacial landscapes. The bedrock (mainly granite and meta-  
114 sediments) underlies glacial drift deposits that cover 60-70% of the catchment and maintain a  
115 perennial source of base flow to the stream (Soulsby, Bradford, et al., 2016). These deposits  
116 are overlain by ~1m deep histosols (peats and peaty gleys) in the riparian area (~21% of the

117 catchment area). The remainder (~79%) of the catchment pedology is dominated by freely-  
118 draining podzols (<0.7 m deep) on the hillslopes, while thin regosols (rankers) are found  
119 where drift deposits are marginal (above 400 masl). Spatial patterns of land cover reflect  
120 these hydrogeological units. Heather shrublands (*Calluna vulgaris* and *Erica* spp.) and Scots  
121 pine (*Pinus sylvestris* L.) are the dominant vegetation over the hillslopes. The former is  
122 secondary vegetation following deforestation and subsequent overgrazing by red deer  
123 (*Cervus elaphus*) and sheep, while Scots pine forest would be the natural vegetation cover,  
124 now restricted to steeper inaccessible hillslopes and fenced plantations. Finally, grasses  
125 (*Molinia caerulea*) cover the riparian gley soils, while the peat is dominated by bog mosses  
126 (*Sphagnum* spp.). The water balance is energy-limited; annual precipitation is ~1000 mm  
127 with a slight winter maximum, about 400 mm becomes evapotranspiration (ET) with  
128 pronounced seasonality (Birkel et al., 2011). Mean annual temperature is 7 °C and monthly-  
129 averaged temperatures remain above 0 °C, the climate is temperate to boreal oceanic; <5% of  
130 precipitation usually occurs as snowfall.

131

## 132 **2.2 Critical Zone ecohydrological model**

133 We used EcH<sub>2</sub>O-iso, a process-based, fully-distributed ecohydrological model designed to  
134 jointly simulate energy, water and vegetation dynamics in the critical zone (Kuppel et al.,  
135 2018a). EcH<sub>2</sub>O-iso tightly couples a two-layer energy balance scheme based on flux-gradient  
136 similarity (allowing to separately compute transpiration, evaporation of intercepted water,  
137 and soil evaporation), a hydrological module for vertical and lateral transfers (based on a 1D  
138 kinematic wave approximation for stream and groundwater), and a biomass component to  
139 simulate vegetation phenology and growth (Maneta & Silverman, 2013). Root water uptake  
140 profile is based on a parameterized exponential form across the three layers of the  
141 hydrological domain (encompassing the vadose zone and groundwater), and soil evaporation  
142 is restricted to the top layer. EcH<sub>2</sub>O-iso tracks stable isotopes (<sup>2</sup>H and <sup>18</sup>O) ratios in water,  
143 and water ages. For each pixel and critical zone compartment, a full-mixing mass balance  
144 equation is applied to tracers (isotopic ratios and age) at each sub-timestep when water is  
145 exchanged (and includes evaporative fractionation of isotopes; Kuppel et al., 2018a),  
146 providing a time-varying mean value. Because outgoing fluxes have the same fully-mixed  
147 tracer signature as their feeding pool(s), the mean water ages (MWA) in the former locally  
148 equates the water ages in the latter. More details on EcH<sub>2</sub>O-iso, and recent developments, can  
149 be found in Smith et al. (2019). The model was run at daily time steps using a 100x100 m<sup>2</sup>  
150 grid, from February 2013 to February 2016 (see section 2.3 and Supplementary Information).

## 151 **2.3 Data sets and model calibration**

152 The model has been extensively calibrated using a wide range of datasets (Kuppel et al.,  
153 2018b, 2018a). Calibration data includes stream discharge at the outlet, soil water content (5  
154 sites, multiple depths), sapflow-derived pine stand transpiration (2 sites), top-of-canopy net  
155 radiation (3 sites), and isotope ratios ( $\delta^2\text{H}$  and  $\delta^{18}\text{O}$ ) at the stream outlet, in bulk soil water (4  
156 sites), and groundwater (4 wells), as summarised in Table S1 and detailed in Kuppel et al.  
157 (2018b, 2018a). This provided an ensemble of 30 “best runs” which satisfactorily captured  
158 catchment behavior in terms of water fluxes, stores, and velocity, as summarized in Fig. S2.  
159 A 12-yr spinup period was used during calibration to limit transient effects in the different  
160 stores, fluxes and isotopic composition. These ensemble simulations were then re-run with a  
161 30-yr spinup, given the longer period necessary for water ages to stabilize (Fig. S3).

## 162 **2.4 Analysis**

163 We exclude canopy-intercepted water from our analysis of water fluxes, stores, and age

164 dynamics. EcH<sub>2</sub>O-iso does not simulate stemflow, so intercepted water does not mix with the  
165 below-canopy compartments; there is field-based evidence that it is a reasonable  
166 simplification as stemflow accounts for <1% of net precipitation here (Soulsby et al., 2017).  
167 We acknowledge that evaporation of intercepted water makes up a substantial part of  
168 evapotranspiration (ET), associated to very young ages (~38±13% of annual ET and aged a  
169 few days, not shown). In the remainder of this study, precipitation inputs are adjusted for  
170 interception losses (effective precipitation, Fig. 1a). Finally, our analysis utilizes the  
171 normalization proposed by (Zuecco et al., 2016), using the maximum and minimum values  
172 for each model realization:

$$x^*(t) = \frac{x(t) - \min(x)}{\max(x) - \min(x)} \quad (1)$$

174

175 where  $x$  is either water age or storage. This allows focusing on the intra-simulation seasonal  
176 dynamics for each of the catchment parameterizations in the ensemble approach adopted  
177 here, removing inter-model dispersion arising from different initial conditions after spinup.

## 178 **3 Results**

### 179 **3.1 Catchment storage and water ages**

180 The below-canopy catchment budget is dominated by stream discharge (Fig. 1b), yet the  
181 relative contribution of soil evaporation and transpiration is dominant during the growing  
182 season (Fig. 1c). A large predictive uncertainty surrounds the age of groundwater outputs  
183 (Fig. 1d), however this flux remains small in all ensemble simulations. As a result, we do not  
184 analyze the patterns of lateral groundwater outflow, and limit our focus to the dynamics of  
185 soil evaporation, plant transpiration, and stream outflow.

186

187 The mean water ages in streamwater and transpiration clearly exhibits an inverse storage  
188 relationship – i.e. water age since precipitation input decreases as catchment storage  
189 (including the unsaturated and saturated profile) increases, *sensu* (Harman, 2015) – across the  
190 simulation ensemble (Fig. 2a,c). However, this inverse relationship is only found during the  
191 growing season for soil evaporation, while water age time tends to increase in winter time  
192 when this flux is small (Fig. 2b). Some anticlockwise hysteresis is also clearly visible for  
193 transpiration, and to a lesser extent soil evaporation (Fig. 2a-b), meaning that for a similar  
194 storage value, these outgoing fluxes tap older water during the drying period (February to  
195 July) than during the rewetting period (August to January).

196

### 197 **3.2 Preference of age extraction**

198 We also computed the ratio between the age of water losses and the age of the water being  
199 currently stored in the catchment. This exit age ratio, hereafter EAR, is indicative of the  
200 “preference” towards younger waters (Soulsby, Birkel, et al., 2016). In other words, EAR is  
201 the degree to which each output mobilizes the youngest stores or rapid catchment flow  
202 pathways (EAR tending to 0, hereby highlighting heterogeneities) to tapping across well-  
203 mixed, less variable stores (EAR tending to 1). We find that EAR values between 0.25 and  
204 0.5 for both green water outputs (Fig. 3a-b). We interpret the limited seasonal variations as a  
205 high synchronicity between catchment water demographics and that of extracted water,  
206 consistent with the full mixing assumptions used at grid cell level and the generally weak  
207 seasonality of catchment inputs and limited water stress in the catchment at any point of the

208 year. By contrast, streamflow shows a preference for comparatively older water pools in the  
209 catchment, while a clear progression towards younger water extraction with increased  
210 wetness states (and high flow), then mobilizing waters half the age (or less) than those stored  
211 in the catchment.

212

### 213 **3.3 Dynamics in dominant hydrogeological units**

214 The above description provides an integrated perspective at the catchment scale of the age  
215 demographics of water losses in the critical zone. However, we also seek to disentangle the  
216 interplay between local dynamics and the relative contributions in different hydrogeological  
217 units. In general, a weaker control of storage on water ages is found in the valley bottom than  
218 on the hillslopes (Fig. 4a-d). In addition, the patterns for transpiration ages at the catchment  
219 scale (Fig. 2a) mostly reflect those found on the hillslopes with an inverse storage effect and  
220 hysteresis (Fig. 4a), related to the dominant contribution of Scots pine and heather shrubs  
221 transpiration covering ~80% of the catchment (in terms of spatially-aggregated budget, Fig.  
222 4e). In the case of soil evaporation (Fig. 4c-d), these individual units displays less  
223 symmetrically V-shaped MWA\* dynamics than found at the catchment scale (Fig. 2a), with a  
224 more limited increase of winter ages on the hillslopes and noisier variations on the valley  
225 bottom. We note that these normalized seasonal dynamics are relative to a baseline of  
226 younger water evapotranspired in the hillslopes than in the valley bottom (< 1 year and 1-2.5  
227 years, respectively, Fig. S4).

## 228 **4 Discussion**

229 The time-variant nature of water ages from precipitation to output has been widely observed  
230 in various critical zone settings (Benettin et al., 2017; Destouni, 1991; Heidbüchel et al.,  
231 2012; Hrachowitz et al., 2015; van der Velde et al., 2012). Taking the example of a well-  
232 studied catchment, we show that while mean water ages are generally controlled by storage,  
233 seasonal patterns are significantly different between the output fluxes and the spatial scales  
234 we considered (catchment and hydrogeological subunits). In particular, our approach reveals  
235 that the “inverse storage effect” defined by Harman (2015) applies to stream water ages at the  
236 catchment scale but is not a ubiquitous feature across outputs and in subcatchment units.

237

238 Water ages in catchment-scale transpiration had a rapid and inverse response to  
239 storage variations particularly visible during the driest months, at the peak of the growing  
240 season. However, the bulk of seasonal dynamics exhibited a hysteretic storage-age behavior  
241 in which spring transpiration used older water than in fall (by ~25% of the seasonal age  
242 range), despite similar catchment storage status (Fig. 2a). This seems to point to a seasonal  
243 carry-over of winter rewetting events that partially subsidize spring transpiration, while plant  
244 water use in fall relies on more recently infiltrated precipitation after summer drying. Such an  
245 asymmetrical buffering effect was less identifiable for soil evaporation where it might be  
246 obscured by short-term variability (Fig. 2b), probably owing to the fact that topsoil dynamics  
247 are more responsive to hydroclimate variability, and thus less prone to “memory effects”,  
248 than the deeper layers accessible to the root water uptake. In addition, the distinctively V-  
249 shaped seasonal storage-age relationship of soil evaporation indicates the fundamentally  
250 different dynamics between the two components of evaporative losses, even at the aggregated  
251 scale first considered here, thereby pointing at the limits of only considering a single  
252 evapotranspiration term to conceptualise ecohydrological couplings driving partitioning and  
253 water budget in catchments (Fatichi et al., 2012, 2014; Vivoni, 2012). Finally, only mean  
254 water ages in streamflow displayed a consistent inverse storage effect (Fig. 2c), a feature  
255 observed in various settings (e.g. Harman, 2015; Pangle et al., 2017; Rodriguez et al., 2018),

256 including the one studied here with independent approaches (Benettin et al., 2017). The  
257 synchronous decrease in the ratio of stream water ages to catchment-stored water ages (Fig.  
258 3c) reflects an increasing relative contribution of recent, less-well-mixed water to the stream  
259 in wetter conditions (Harman, 2019). In other words, during the rewetting period (August to  
260 January), stream-feeding pathways increasingly bypass and outpace contribution of older,  
261 mixed water in deeper catchment stores. The associated seasonal decrease of transit times  
262 (since lateral and precipitation inputs) in the valley bottom, but also on the hillslopes in the  
263 wettest months (Fig. S5c) is consistent with the reported increase in extent of surface  
264 connectivity via saturation overland flow, first in the saturated riparian histosols and  
265 eventually including the hillslope podzols (Tetzlaff et al., 2014).

266

267 In the valley bottom, the weak seasonal variability of water ages (and in particular the  
268 lack of water ages increase in summer, Fig. 4) and markedly older baseline ages than the  
269 hillslopes (Fig. S4) translated the generally wet conditions in the riparian area, maintaining  
270 leveled soil evaporation rate throughout the months when storage is lowest (Fig. 4f and S4e).  
271 This is a result of slow-draining water subsidy from the hillslope with downslope transit time  
272 of about 6 months and even slower turnover within the valley bottom with 1-to-2 years transit  
273 times (Fig. S5a-b). The stronger (and generally inverse) constraint of storage on MWA on the  
274 hillslope reflected the higher turnover in these shallower freely-draining podzols, where the  
275 limited hydroclimatic seasonality makes green water demographics more controlled by the  
276 atmospheric demand.

277

278 These differences between subcatchment units draw attention to the interplay between  
279 local dynamics of water ages and spatial organization of contributing fluxes, whereby a range  
280 of catchment age dynamics that could be interpreted as complex mixing patterns may also be  
281 produced by time-changing lateral heterogeneity. Using the daily fraction of catchment  
282 transpiration and soil evaporation budgets taking place on the hillslope as indices of such an  
283 heterogeneity, we found that landscape organization explains around half of the catchment-  
284 scale, seasonal variability of the ages of these two green water outputs ( $43\pm 25\%$  and  $53\pm 20\%$   
285 across ensemble simulations, respectively, not shown). Although our numerical model uses  
286 full local mixing in each simulated compartment for each pixel, our spatially-aggregated  
287 analyses of storage and water ages essentially describe a partially-mixed system with  
288 changing macro-scale water pathways over time. Our approach thus provides an alternative to  
289 catchment-lumped analytical formulations of time-variant transit time distributions (TTD)  
290 and StorAge selection (SAS) functions and to numerical experiments testing the impact of  
291 partially-mixing flows on water ages (e.g., Cain et al., 2019; Knighton et al., 2017). All  
292 provide different perspectives on how to model the unavoidable structural and functional  
293 heterogeneity of real-world catchments (Sprenger et al., 2019), but spatially-lumped  
294 descriptions cannot fully reject the full mixing hypothesis if a complex age response can also  
295 be reproduced with spatial heterogeneity, as is the case here.

296

297 There is thus a need to assess the relative importance of preferential flows at the pore  
298 scale (Beven & Germann, 2013) and areal scales (i.e. spatial organization; Hendrickx &  
299 Flury, 2001) in the resulting catchment functioning. Our approach was limited to considering  
300 uniform flows in three hydrological layers of each grid cell, and our soil evaporation ages  
301 from a well-mixed topsoil likely underestimated the contribution of recent surface waters to  
302 this flux. Yet, combining these simplified formulations with spatially-distributed flow paths  
303 has allowed capturing water flux and tracer concentrations measured at multiple locations in  
304 several critical zone compartments of high-latitude catchments (Kuppel et al., 2018a; Smith  
305 et al., 2019). In addition, accounting for local preferential flow may not always improve

306 catchment-scale simulations (Glaser et al., 2019; Hopp et al., 2020). The present simulations  
307 further suggest that plants access relatively old water pools (from ~6 months on the  
308 hillslopes, to over 2 years in the valley bottom, Fig. S4) with an hysteretic relationship to  
309 storage not found in streamflow ages. This modelling approach is therefore relevant to  
310 interpreting water pathways from measured tracer concentrations (Penna et al., 2018). It can  
311 help evaluating whether the proposed ecohydrological separation (Brooks et al., 2010)  
312 between tightly-bound and mobile water effectively reflects “preferential pore space  
313 selection” in root water uptake, or can be fully explained via the conservation of water  
314 masses in heterogeneously-conducting media (Berghuijs & Allen, 2019) including tree-water  
315 storage (Knighton et al., 2020; Meinzer et al., 2006).

316 The generic nature of hillslope-riparian couplings in real-world catchments makes our  
317 spatialized analysis of water ages relevant to identifying key contributing areas in the face of  
318 environmental changes. We show for example that hillslope transit times account for a third  
319 to half of streamflow ages at the outlet, suggesting high sensitivity of river flow to water  
320 partitioning on hillslopes. Yet, the vast majority of catchment management solutions focus on  
321 riparian management (e.g. forest felling restrictions zones around streams, or water quality  
322 buffer strips in agricultural land) as common-sense or cost-effective approaches, without  
323 often knowing the specific nature of hillslope-riparian interactions. These issues may be  
324 particularly crucial in high-latitude landscapes such as the one where this study was  
325 conducted. There, climate and land cover changes induce rapid shifts in snowfall/rainfall  
326 partition and biomes shifts. The presented ecohydrological modelling approach may help  
327 inferring the likely ecohydrological consequences of these changes, more critically so  
328 considering the general decline of high-latitude catchment monitoring (Laudon et al., 2017).

329

### 330 **Acknowledgments, model code, and data**

331 The source code of the EcH<sub>2</sub>O-iso model is publicly available on  
332 [https://bitbucket.org/sylka/ech2o\\_iso/](https://bitbucket.org/sylka/ech2o_iso/) (branch master\_2.0). Input, forcing and output files,  
333 along with scripts to launch ensemble simulations and create basic plots are on a Zenodo  
334 repository (doi: 10.5281/zenodo.3592491). This work was supported by the European  
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517

518

519 **Figure 1.** (a) Daily catchment-aggregated time series of the below-canopy water budget at  
520 Bruntland Burn. Outputs are decomposed into soil evaporation, transpiration, stream and  
521 groundwater discharge –note the right y-axis for stream discharge– (b), showing their relative  
522 contributions to below-canopy outputs (c), and mean water ages (d-e). Errors bars (a) and  
523 ribbons (a-e) show the 80% intervals across the simulation ensemble.

524

525 **Figure 2.** (a-c) Spatially-aggregated plots of normalized mean water ages (MWA\*) in plant  
526 transpiration(a), soil evaporation (b), and outlet stream discharge (c) versus normalized  
527 catchment storage ( $S^*_{\text{catchment}}$ ). For each day of the year (color scale), multi-year means  
528 (02/2013 – 02/2016 period) are shown using cross-ensemble medians (small points) and 50%  
529 confidence ellipses. Open triangles indicate the normalized median storage value for each  
530 simulation. (d) Mean seasonal cycle of each flux contribution to the below-canopy outputs  
531 (80% interval across the simulation ensemble).

532

533 **Figure 3.** Spatially-aggregated plots of exit age ratio (EAR, the ratio between the age of  
534 water in outputs and that of water currently stored in the critical zone) of plant transpiration  
535 (a), soil evaporation (b), and outlet stream discharge (c) versus the normalized catchment  
536 storage ( $S^*_{\text{catchment}}$ ). For each day of the year (color scale), multi-year means (02/2013 –  
537 02/2016 period) are shown using cross-ensemble medians (small points) and 50% confidence  
538 ellipses. Open triangles indicate the normalized median storage value for each simulation.

539

540 **Figure 4.** (a-b) Spatially-aggregated plots of normalized mean water ages (MWA\*) of  
541 transpiration against the normalized storage ( $S^*$ ) state of the subsurface, both spatially  
542 aggregated over two sub-catchment units: the hillslopes overlain by podzolic and ranker soils  
543 (a, ~80% of the area), and the valley bottom overlain by peat/peaty gley soils (b, ~20% of the  
544 area). (c-d) Same as (a-b) with the MWA\* of soil evaporation. For each day of the year  
545 (color scale), multi-year means (02/2013 – 02/2016 period) are shown using cross-ensemble  
546 medians (small points) and 50% confidence ellipses. Open triangles indicate the normalized  
547 median storage value for each simulation. (e-f) Mean seasonal cycle of the contributions of  
548 soil evaporation (e) and plant transpiration (f) to catchment-scale below-canopy outputs (80%  
549 interval across the simulation ensemble); note the difference of vertical scale between panel  
550 (e) and (f).

551

Figure 1.

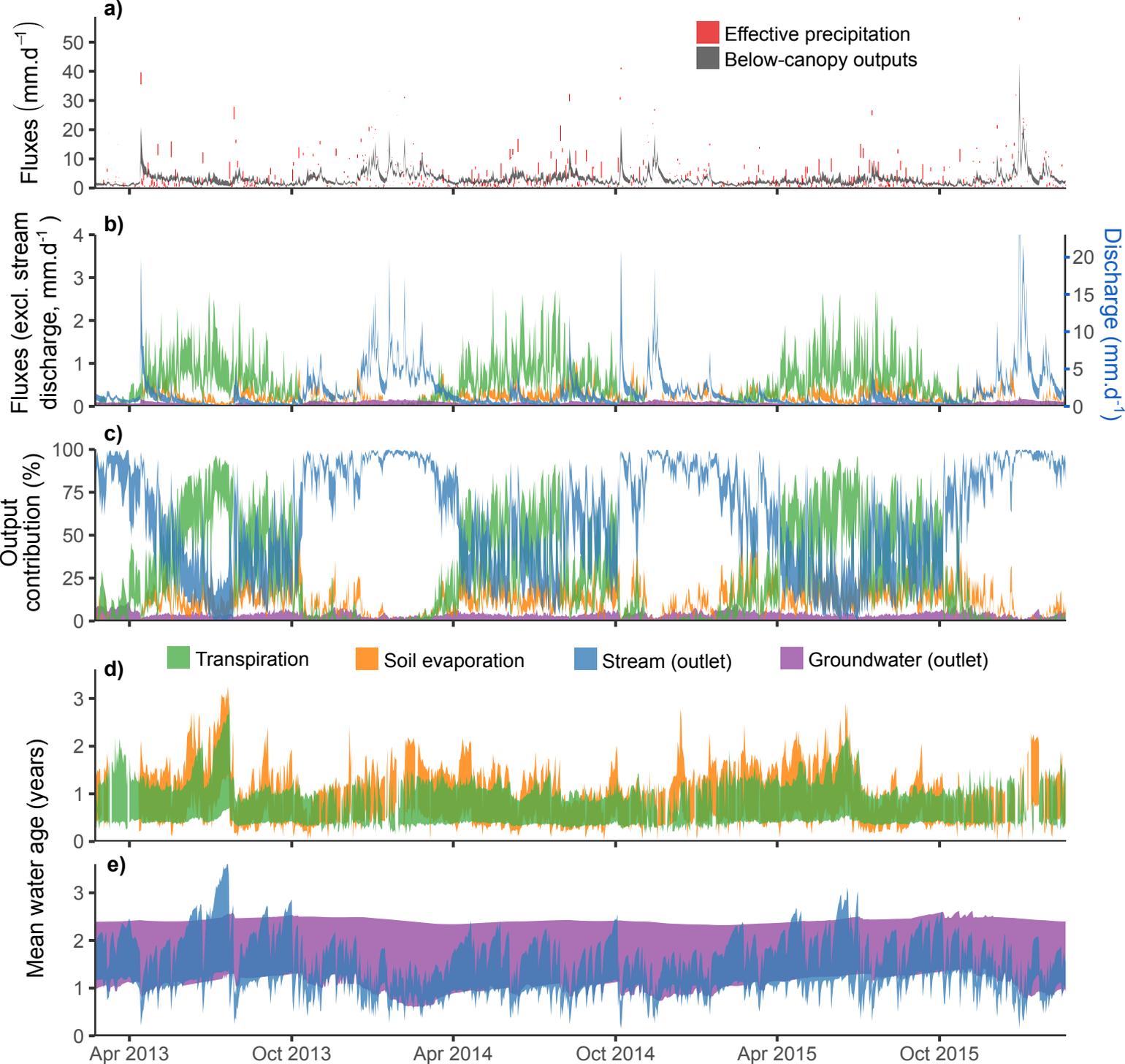


Figure 2.

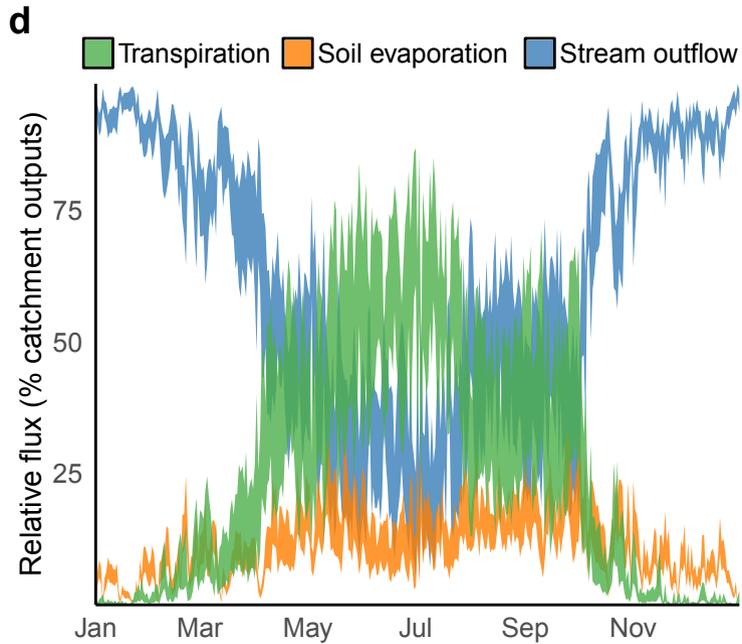
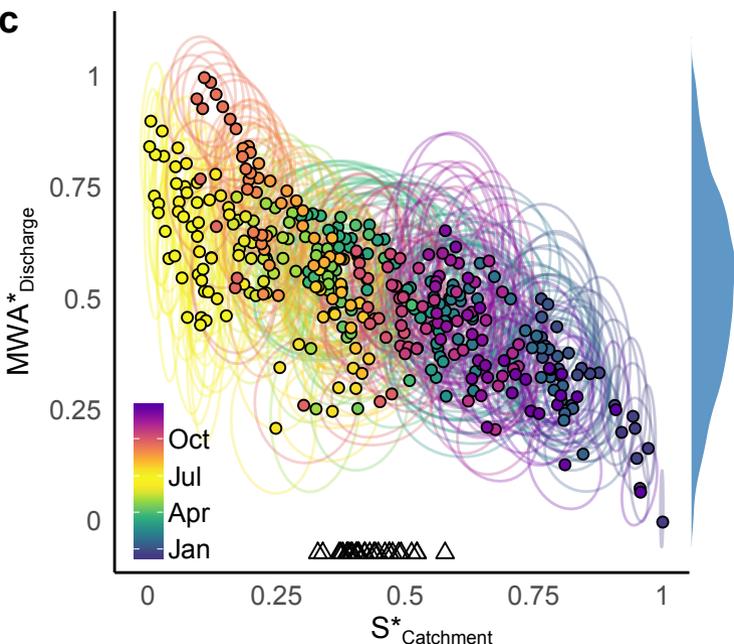
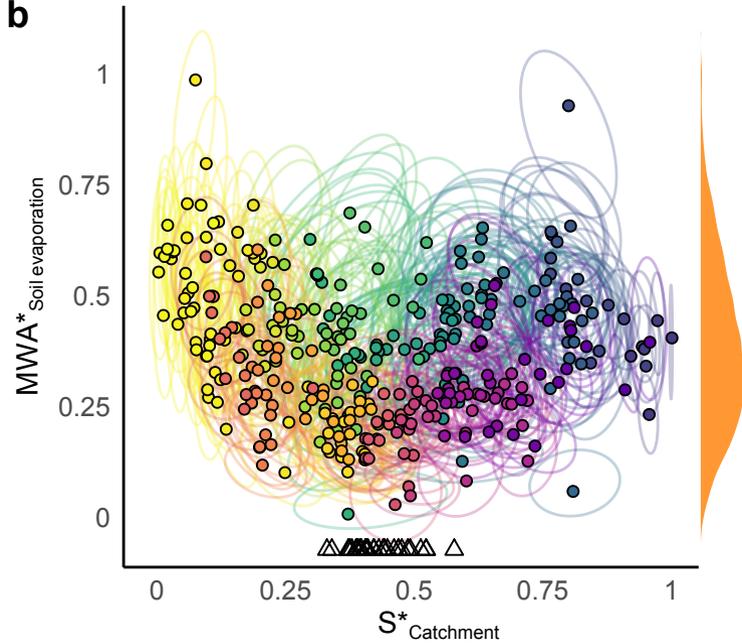
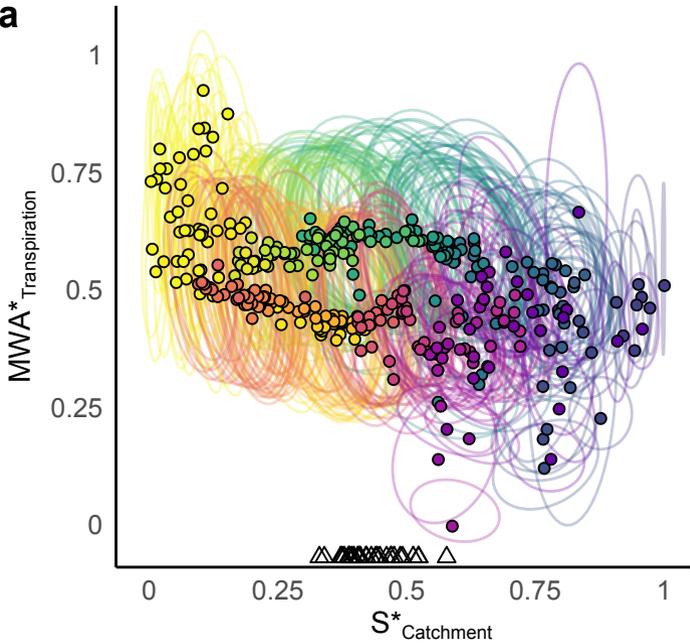


Figure 3.



**Figure 4.**

